

## Thermodynamics of Cu<sup>2+</sup> Adsorption on soil Humin

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Received 15 June 2014;

Revised 30 Aug. 2014;

Accepted 4 Nov. 2014

**ABSTRACT:** The adsorption thermodynamic characteristic of Cu<sup>2+</sup> from aqueous solution onto humin and also humic acid from a typical black soil in northeast China were examined at three different temperatures (298, 318 and 338 K) by batch isotherm experiments. Results showed that humin was structurally different from humic acid. The increase of temperature had a positive effect on the adsorption process. Freundlich equation described the equilibrium data better with respect to Langmuir equation. Thermodynamic parameters, i.e. standard free energy changes ( $\Delta G^\circ$ ), standard enthalpy change ( $\Delta H^\circ$ ) and standard entropy change ( $\Delta S^\circ$ ), revealed that Cu<sup>2+</sup> adsorption by humin and humic acid were a feasible, spontaneous and endothermic process with an increase in disorder. The values of isosteric heat of adsorption ( $\Delta H_x$ ) increased with increasing surface Cu<sup>2+</sup> loading, and the increase in  $\Delta H_x$  values was smaller for humin than for humic acid. The results indicated that humin and humic acid possessed energetically heterogeneous surfaces. Humic acid had greater surface heterogeneity than humin

**Key words:** Humin, Humic acid, Cu<sup>2+</sup>, Adsorption isotherm, Thermodynamic parameters

## INTRODUCTION

Soil organic matter is one of the most important soil components. It has a profound effect on the physical, chemical and biological functions of soil, including aggregate stability, water-holding capacity, cation exchange capacity, nutrient retention and release, etc. (Krull *et al.*, 2004). Humic substances represent the largest constituent (60-80%) of soil organic matter, responsible for many complex chemical reactions in soil such as adsorption, ion exchange, complexation, redox, dissolution, and precipitate (Stevenson, 1994; Schnitzer, 1999). Generally, humic substances can be divided into three main fractions based on the basis of their solubility in aqueous solution at different pH values: humic acid, which is not soluble in water at pH<2 but is soluble at less acidic pH values; fulvic acid, which is soluble in water at all pH values; and humin, which is not soluble in water at any pH value (Aiken *et al.*, 1985; Stevenson, 1994). One of the most striking features of humic substances is their strong interaction with inorganic and organic pollutants (Perminova *et al.*, 2005). The adsorption of metal ions on humic substances has long

been known to play an important role in controlling the behavior and fate of trace metals in the environment. The majority of studies published to date have focused on the adsorption characteristics of metal ions on base-soluble fractions of humic substances, namely humic and fulvic acids. However, very little information is available on humin. Only a few have reported on the adsorption properties of metal ions on humin from peat (Contreras *et al.*, 2006; De la Rosa *et al.*, 2003; Gardea-Torresdey *et al.*, 1996), lignite (Havelcová *et al.*, 2009), and brown coal (Alvarez-Puebla *et al.*, 2005; Alvarez-Puebla *et al.*, 2006). In addition, Helal *et al.* (1998, 2006a, 2006b) and Rigol *et al.* (1998) reported the adsorption features of some radionuclides (<sup>134</sup>Cs, <sup>137</sup>Cs, <sup>90</sup>Sr, <sup>153</sup>Gd, <sup>110m</sup>Ag, <sup>60</sup>Co and <sup>65</sup>Zn) on humin from peat soil or agricultural soil.

Copper (Cu) is an essential trace element for plant nutrition, while it could become toxic when its concentration is high in the soil. In our previous study, we determined, for the first time, the adsorption isotherm of Cu<sup>2+</sup> on humin isolated from a forest soil

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using batch equilibration method (Zhang *et al.*, 2012). However, the thermodynamic properties of the adsorption of  $\text{Cu}^{2+}$  on soil humin, which can provide useful information for evaluating the mechanisms of adsorption process, are still lacking. On the other hand, the interaction of humin with metal ions is closely related to their chemical structures. Elemental analysis and solid-state  $^{13}\text{C}$  cross-polarization magic angle spinning nuclear magnetic resonance ( $^{13}\text{C}$  CPMAS NMR) spectroscopy are powerful tools for elucidating the structural characteristics of humin (Rice and MacCarthy, 1991; Simpson *et al.*, 2011). The black soil is important for its high level of fertility and productivity. Black soil region in northeast China is one of the three major black lands in the world. The main objective of the present study are to evaluate the thermodynamics of  $\text{Cu}^{2+}$  adsorption on humin from a typical black soil in northeast China at different temperatures (298, 318 and 338 K) by conducting batch isotherm experiments. Moreover, the isolated humin are characterized using elemental analysis and solid-state  $^{13}\text{C}$  CPMAS NMR spectroscopy. The differences of structural and  $\text{Cu}^{2+}$  adsorption characteristics between humin and humic acid are also compared.

## MATERIALS & METHODS

The soil used in the study is black soil (Haplic haezem according to the FAO Soil Classification, and Typic Hapludoll according to the USDA Soil Taxonomy). The surface soil (0-20 cm) samples were collected from a cultivated maize field located at Changchun City (43°48' 443 N, 125°23' 453 E, 230.7 m above sea level), Jilin Province, Northeast China. The collected soil samples were air-dried, milled and sieved through a 2 mm sieve. Selected physical and chemical properties of the soil were analyzed using the procedure recommended (Lao, 1988), and the results are presented in Table 1. The extraction, fractionation and purification of humic acid and humin were performed following the procedure described by Zhang *et al.* (2009b, 2011a). Briefly, 100 g of each soil sample was first suspended in distilled water and dilute HCl to remove poorly decomposed light fractions and carbonates, respectively. The soil residue was then extracted with a solution of 0.1 mol L<sup>-1</sup> NaOH and 0.1 mol L<sup>-1</sup> Na<sub>4</sub>P<sub>2</sub>O<sub>7</sub> using a 1:10 sample:extractant ratio at room temperature. The extraction procedure was repeated 25 times until the supernatant was almost colorless. The combined

alkaline supernatant was acidified to pH 1.0 with 6 mol L<sup>-1</sup> HCl to separate the humic acid fraction. The insoluble solid residue remaining after alkaline extraction was referred to as humin fraction. The humic acid fraction was purified by de-ashing with 0.5% HCl-HF solution after three cycles of dissolution in 0.1 mol L<sup>-1</sup> NaOH and reprecipitation with 6 mol L<sup>-1</sup> HCl. The humin fraction was purified by de-ashing with 20% HCl-HF solution. The humic acid and humin samples were then further purified by dialyzing against distilled water until they were Cl<sup>-</sup>-free. Finally, the purified humic acid and humin were freeze-dried, ground to <0.1 mm and stored in a desiccator for further use.

The C, H, N, and S contents were determined using an Elementar Vario MICRO elemental analyzer (Germany), and the O content was calculated by mass difference. The ash contents were measured by thermogravimetric analysis performed with a Shimadzu DTG-60 thermal analyzer (Japan) from room temperature to 750 °C at a heating rate of 10 °C min<sup>-1</sup>. Solid-state  $^{13}\text{C}$  cross-polarization magic angle spinning nuclear magnetic resonance ( $^{13}\text{C}$  CPMAS NMR) spectra were obtained using a Bruker AVANCE III 400 WB spectrometer (Switzerland) operating at 100.6 MHz. The operating conditions were as follows: spinning rate 12 kHz, contact time 2 ms, and recycle time 6 s. The spectra were divided into four main chemical shift regions: alkyl C (0-50 ppm), O-alkyl C (50-110 ppm), aromatic C (110-160 ppm), and carbonyl C (160-200 ppm). The total range of O-alkyl C was further divided into regions of methoxyl C (50-60 ppm) and carbohydrate C (60-110 ppm), and the total range of aromatic C into regions of aryl C (110-145 ppm) and phenolic C (145-160 ppm) (Zhang *et al.*, 2011a). The relative intensity of each chemical shift region was determined using the integration routine of the spectrometer. The correction of spinning side band (SSB) was made according to Conte *et al.* (1997).

The adsorption experiments were carried out under ambient conditions by using batch technique. The solid adsorbents (humic acid and humin), NaNO<sub>3</sub> background electrolyte solution and Cu(NO<sub>3</sub>)<sub>2</sub> stock solution were added into 50 mL polyethylene centrifuge tubes to achieve the desired concentrations of different components (i.e. adsorbent dosage 0.2 g L<sup>-1</sup>, ion strength 0.01 mol L<sup>-1</sup> NaNO<sub>3</sub>, initial Cu<sup>2+</sup> concentration 0-80 mg L<sup>-1</sup>). The initial pH of the solution was adjusted to 5.5 with 0.1 mol L<sup>-1</sup> HNO<sub>3</sub> and 0.1 mol L<sup>-1</sup> NaOH, but no further

**Table 1. Selected properties of the soil used in the experiment**

Organic C (g kg <sup>-1</sup> )	Total N (g kg <sup>-1</sup> )	Total P (g kg <sup>-1</sup> )	pH	Sand (20-2000 μm) (g kg <sup>-1</sup> )	Silt (2-20 μm) (g kg <sup>-1</sup> )	Clay (<2 μm) (g kg <sup>-1</sup> )
11.2	1.33	0.47	6.50	472.3	256.2	271.5

pH adjustment was conducted during the adsorption process. The suspensions were shaken at a constant speed of 140 rpm in a shaking water bath with temperature 298, 318 and 338 K, respectively. After shaking the tubes for 24 h, the solid and liquid phases were separated by centrifugation at 16000 rpm for 15 min. The solution was then filtered, and the filtrate was analyzed for Cu<sup>2+</sup> concentration using TAS-990 atomic absorption spectrometry (China). All experimental data were the average of duplicate or triplicate determination.

The amounts of adsorbed Cu<sup>2+</sup> were calculated by the mass balance equation [Eq. (1)]:

$$q_e = (C_0 - C_e) \frac{V}{m} \quad (1)$$

where  $q_e$  is the amount of Cu<sup>2+</sup> adsorbed at equilibrium (mg g<sup>-1</sup>),  $C_0$  is the initial Cu<sup>2+</sup> concentration (mg L<sup>-1</sup>),  $C_e$  is the equilibrium Cu<sup>2+</sup> concentration in solution (mg L<sup>-1</sup>),  $V$  is the volume of the solution (L),  $m$  is the mass of the adsorbents (g).

The adsorption isotherm data were fitted using the Langmuir model [Eqs. (2) and (3)] and the Freundlich model [Eqs. (4) and (5)]:

$$q_e = \frac{k_1 q_m C_e}{1 + k_1 C_e} \quad (\text{non-linear form}) \quad (2)$$

$$\frac{C_e}{q_e} = \frac{C_e}{q_m} + \frac{1}{q_m k_1} \quad (\text{linear form}) \quad (3)$$

$$q_e = k_2 C_e^{1/n} \quad (\text{non-linear form}) \quad (4)$$

$$\ln q_e = \frac{1}{n} \ln C_e + \ln k_2 \quad (\text{linear form}) \quad (5)$$

where  $q_m$  and  $k_1$  are the Langmuir constants related to the maximum adsorption capacity (mg g<sup>-1</sup>) and adsorption energy (L mg<sup>-1</sup>), respectively;  $k_2$  and  $n$  are the Freundlich constants represent the adsorption capacity (mg g<sup>-1</sup> (mg L<sup>-1</sup>)<sup>-1/n</sup>) and adsorption intensity, respectively.

The thermodynamic parameters of adsorption, i.e., the standard free energy changes ( $\Delta G^\circ$ , kJ mol<sup>-1</sup>), the standard enthalpy change ( $\Delta H^\circ$ , kJ mol<sup>-1</sup>) and the standard entropy change ( $\Delta S^\circ$ , J mol<sup>-1</sup> K<sup>-1</sup>), were

calculated from the temperature dependent adsorption isotherms using classical thermodynamic equations [Eqs. (6)-(8)] (Mahmoodi *et al.*, 2011; Sheng *et al.*, 2009):

$$K_d = \frac{C_0 - C_e}{C_e} \frac{V}{m} \quad (6)$$

$$\ln K_d = \frac{\Delta S^\circ}{R} - \frac{\Delta H^\circ}{RT} \quad (7)$$

$$\Delta G^\circ = \Delta H^\circ - T\Delta S^\circ \quad (8)$$

where  $K_d$  is the thermodynamic equilibrium constant (mL g<sup>-1</sup>),  $R$  is the ideal gas constant (8.314 J mol<sup>-1</sup> K<sup>-1</sup>),  $T$  is the reaction temperature (K). The values of  $\Delta H^\circ$  and  $\Delta S^\circ$  are obtained from the slope and y-intercept of the plot of  $\ln K_d$  versus  $1/T$ .

The isosteric heat of adsorption ( $\Delta H_x$ , kJ mol<sup>-1</sup>) at constant surface loading ( $q_e=2, 4, 6, 8, 10, 12, 14, 16$  mg g<sup>-1</sup>) was calculated using the Clausius-Clapeyron equation [Eq. (9)] (Anirudhan and Radhakrishnan, 2011):

$$\frac{d(\ln C_e)}{dT} = -\frac{\Delta H_x}{RT^2} \quad (9)$$

where  $C_e$  (mg L<sup>-1</sup>) is the equilibrium Cu<sup>2+</sup> concentration at constant  $q_e$  (mg g<sup>-1</sup>), which was obtained from the adsorption isotherm data at various temperatures  $T$  (K). The  $\Delta H_x$  values were calculated from the slope resulting from plotting of  $\ln C_e$  versus  $1/T$ .

## RESULTS & DISCUSSION

The elemental composition and atomic ratios of humic acid and humin are presented in Table 2. Compared with humic acid, humin contained more C, H and S and less N and O. The H/C ratio was lower whereas the O/C ratio was higher for humic acid than for humin, indicating that humin was more aliphatic and less polar (Xing *et al.*, 2005). Meanwhile, the C/N ratio of humic acid was lower than that of humin, suggesting that resistant forms of N were depleted in humin during the humification process (Zhang *et al.*, 2011b).

The <sup>13</sup>C CPMAS NMR spectra of humic acid and humin are shown in Fig. 1. The spectra exhibited major

**Table 2. Elemental composition and atomic ratios of humic acid and humin from black soil**

Samples	C <sup>a</sup> (g kg <sup>-1</sup> )	H <sup>a</sup> (g kg <sup>-1</sup> )	N <sup>a</sup> (g kg <sup>-1</sup> )	S <sup>a</sup> (g kg <sup>-1</sup> )	O <sup>a</sup> (g kg <sup>-1</sup> )	H/C (atomic ratio)	O/C (atomic ratio)	C/N (atomic ratio)
Humic acid	575.4	39.0	37.0	4.11	344.5	0.813	0.449	18.1
Humin	638.1	52.1	19.1	5.13	285.2	0.981	0.335	38.3

<sup>a</sup> On a ash-free basis.

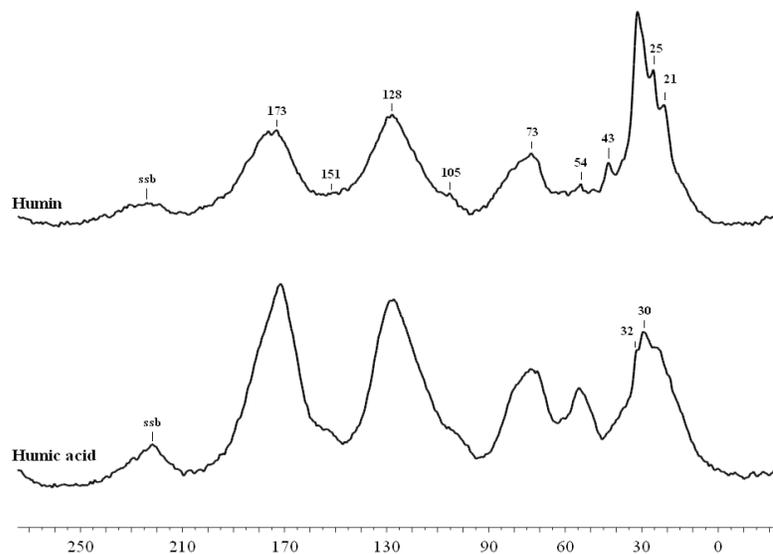
**Table 3. Relative carbon distribution (%) in different regions of chemical shift in <sup>13</sup>C CPMAS NMR spectra of humic acid and humin from black soil**

Samples	Chemical shift regions (ppm)								A/ O-A <sup>a</sup>	Alip/ Arom <sup>b</sup>	HB/ HI <sup>c</sup>
	0-50	50-60	60-110	0-110	110-145	145-160	110-160	160-200			
Humic acid	21.9	6.45	18.5	46.9	23.2	3.23	26.5	26.9	0.88	1.77	0.93
Humin	34.7	2.57	17.5	54.8	19.8	2.37	22.2	23.1	1.74	2.47	1.32

<sup>a</sup> alkyl C/O-alkyl C = (0-50)/(50-110).

<sup>b</sup> aliphatic C/aromatic C = [(0-50)+(50-110)]/(110-160).

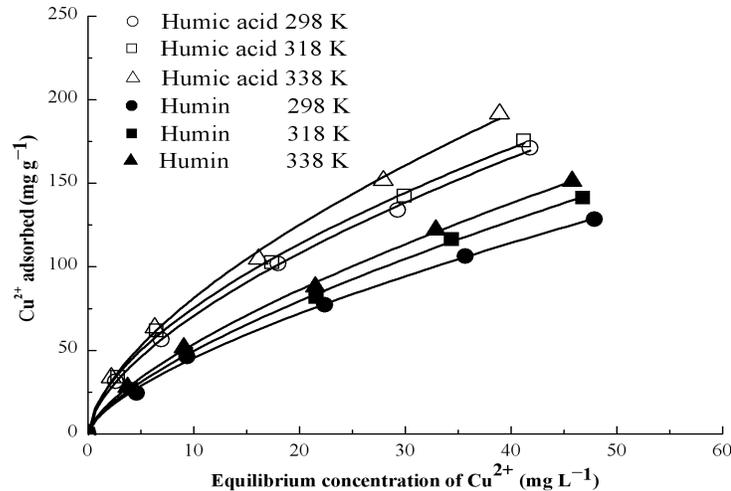
<sup>c</sup> hydrophobic C/hydrophilic C = [(0-50)+(110-160)]/[(50-110)+(160-200)].

**Fig. 1. <sup>13</sup>C CPMAS NMR spectra of humic acid and humin from black soil**

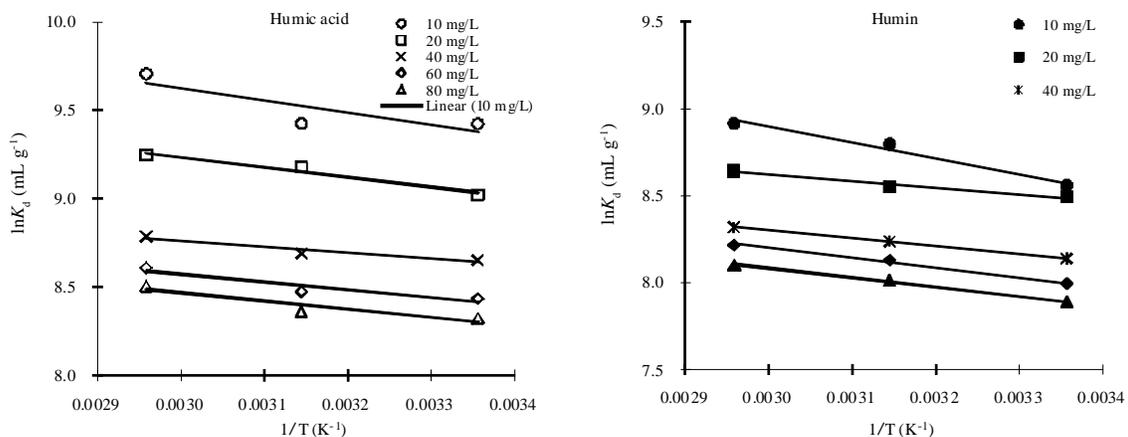
signal peaks at 21, 25, 30, 32, 43, 54, 73, 105, 128, 151 and 173 ppm. Based on Zhang *et al.* (2009b; 2011a), the peaks at 21, 25, 30, 32 and 43 ppm in the alkyl C region were assigned as -CH<sub>3</sub>-, -CH<sub>2</sub>-, amorphous -(CH<sub>2</sub>)<sub>n</sub>-, crystalline -(CH<sub>2</sub>)<sub>n</sub>-, and branched aliphatic C, respectively. The signals at 54, 73 and 105 ppm in the O-alkyl C region were ascribed to methoxyl C in lignin, and -CHOH-, and anomeric C in carbohydrate, respectively. The signals at 128 and 151 ppm in the aromatic C region represented C- and H-substituted, and O-substituted aromatic C from lignin, respectively. The signal at 173 ppm in the carbonyl C region was indicative of carboxylic acid, amide and ester. Moreover, a distinct peak appeared in the range of 222–224 ppm was assigned as spinning side band (SSB).

The relative intensities of C functional groups of humic acid and humin are presented in Table 3. Across the two humic substance fractions, the intensity of aliphatic C (the sum of alkyl C and O-alkyl C, 46.9%-54.8%) was higher than that of aromatic C (22.2%-26.5%) and carbonyl C (23.1%-26.9%). Moreover,

carbohydrate C showed the highest intensity (17.5%-18.5%) in the O-alkyl C region, and the intensity of aryl C (19.8%-23.2%) was the highest in the aromatic C region. Compared with humic acid, humin contained larger proportions of alkyl C, but smaller proportions of methoxyl C, carbohydrate C, aryl C, phenolic C and carbonyl C. The ratios of alkyl C/O-alkyl C, aliphatic C/aromatic C and hydrophobic C/hydrophilic C of humin were higher than those of humic acid. The ratios of alkyl C/O-alkyl C, aliphatic C/aromatic C and hydrophobic C/hydrophilic C have been used as indices of the degrees of humification, aliphaticity and hydrophobicity of humic substances, respectively. Larger values of the ratios indicate that humic substances are more humified, aliphatic and hydrophobic (Zhang *et al.*, 2011a). Furthermore, alkyl C/O-alkyl C ratio is generally considered a more suitable index for describing the humification degree of humic substance fractions with respect to other indices such as aromaticity (Chen and Chiu, 2003; Mathers and Xu, 2003). Thus, our results implied that humin was the more humified, aliphatic and hydrophobic compared



**Fig. 2.**  $\text{Cu}^{2+}$  adsorption isotherms of humic acid and humin from black soil at different temperatures. Symbols denote experimental data, and lines represent the model fitting of Freundlich equation



**Fig. 3.** Linear plot of  $\ln K_d$  vs.  $1/T$  for  $\text{Cu}^{2+}$  adsorption on humic acid and humin from black soil

with corresponding humic acid. The result of aliphatic C/aromatic C ratio was in accordance with that of H/C ratio.

The adsorption isotherms of  $\text{Cu}^{2+}$  on humic acid and humin at three different temperatures are shown in Fig. 2. It could be seen that the amounts of adsorbed  $\text{Cu}^{2+}$  on the two humic substance fractions increased with increasing  $\text{Cu}^{2+}$  concentration. On the other hand, the amounts of  $\text{Cu}^{2+}$  adsorbed on the two adsorbents also increased with rising temperature, indicating that the adsorption of  $\text{Cu}^{2+}$  was promoted at higher temperature. The possible interpretation for the promotion of  $\text{Cu}^{2+}$  adsorption at higher temperature is that the adsorption of well hydrated  $\text{Cu}^{2+}$  on humic acid and humin is an endothermic process in essence (Li *et al.*, 2011; Sheng *et al.*, 2009). Compared with humic acid, the adsorption

amounts of humin for  $\text{Cu}^{2+}$  were lower. The adsorption percentages of  $\text{Cu}^{2+}$  ranged from 47.7% to 79.2% for humic acid and from 40.1% to 62.5% for humin. The parameters of linearized Langmuir and Freundlich equations obtained by fitting the isotherms are given in Table 4. The adsorption data could be well described by both Langmuir and Freundlich equations with coefficient of determination ( $R^2$ ) greater than 0.925. However, the Freundlich equation gave a better fit with  $R^2$  values ranging from 0.993 to 1.000. Based on the Langmuir equation, the maximum adsorption capacities of  $\text{Cu}^{2+}$  ( $q_m$ ) on humic acid and humin were 243.9 and 227.3  $\text{mg g}^{-1}$  at 298 K, respectively. The higher adsorption capacities of humic acid to  $\text{Cu}^{2+}$  was confirmed by the larger values of the Langmuir constant  $k_1$  and the Freundlich constant  $k_2$  for humic acid than for humin. The  $n$  values in the Freundlich equation were larger than 1, indicating a favorable and

**Table 4. Adsorption isotherm parameters derived from linear Langmuir and Freundlich equations for Cu<sup>2+</sup> adsorption on humic acid and humin from black soil at three different temperatures**

Samples	T (K)	Langmuir			Freundlich		
		$q_m^a$	$k_1^b$	$R^{2e}$	$n^c$	$k_2^d$	$R^{2e}$
Humic acid							
	298	243.9	0.047	0.953	1.667	18.07	1.000
	318	250.0	0.050	0.962	1.699	19.64	0.997
	338	270.3	0.052	0.925	1.718	22.00	0.998
Humin							
	298	227.3	0.025	0.975	1.433	8.791	0.993
	318	243.9	0.027	0.952	1.475	10.43	0.999
	338	256.4	0.028	0.946	1.488	11.51	0.999

<sup>a</sup>  $q_m$ : Maximum adsorption quantity (mg g<sup>-1</sup>). - <sup>b</sup>  $k_1$ : Langmuir constants (L mg<sup>-1</sup>). - <sup>c</sup>  $n$ : Freundlich constants (dimensionless).  
<sup>d</sup>  $k_2$ : Freundlich constants (mg g<sup>-1</sup> (mg L<sup>-1</sup>)<sup>-1/n</sup>). - <sup>e</sup>  $R^2$ : Coefficient of determination.

**Table 5. Values of thermodynamic parameters for the adsorption of Cu<sup>2+</sup> on humic acid and humin from black soil.**

Samples	$C_0$ (mg L <sup>-1</sup> )	$\Delta H^\circ$ (kJ mol <sup>-1</sup> )	$\Delta S^\circ$ (J mol <sup>-1</sup> K <sup>-1</sup> )	$\Delta G^\circ$ (kJ mol <sup>-1</sup> )		
				298 K	318 K	338 K
Humic acid						
	10	5.80	97.4	-23.2	-25.2	-27.1
	20	4.70	90.0	-22.4	-24.2	-26.0
	40	2.81	81.3	-21.4	-23.0	-24.7
	60	3.60	82.1	-20.9	-22.5	-24.1
	80	3.81	81.8	-20.6	-22.2	-23.8
Humin						
	10	7.50	96.4	-21.2	-23.2	-25.1
	20	3.06	80.9	-21.0	-22.6	-24.3
	40	3.71	80.1	-20.2	-21.8	-23.4
	60	4.68	82.2	-19.8	-21.5	-23.1
	80	4.43	80.5	-19.6	-21.2	-22.8

heterogeneous adsorption of Cu<sup>2+</sup> on the two fractions of humic substances (Zhang *et al.*, 2009a). The degree of adsorption favorability and heterogeneity was higher for humic acid than for humin, which was also consistent with the larger Cu<sup>2+</sup> adsorption of the former. The greater amounts of Cu<sup>2+</sup> adsorption on humic acid than on humin could be attributed to the higher carboxyl and phenolic hydroxyl groups of the former (Table 3). On the other hand, the Langmuir and Freundlich isotherm parameters all increased with the increase in temperature, which agreed with the promotion of Cu<sup>2+</sup> adsorption at higher temperature. The thermodynamic parameters calculated from Eqs. (6) to (8) are listed in Table 6, and the relevant plot of  $\ln K_d$  vs.  $1/T$  is given in Fig. 3. The magnitude of  $\Delta G^\circ$ ,  $\Delta H^\circ$ , and  $\Delta S^\circ$  values ranged from -20.6 kJ mol<sup>-1</sup>, 3.81 kJ mol<sup>-1</sup>, 81.3 J mol<sup>-1</sup> K<sup>-1</sup> to -27.1 kJ mol<sup>-1</sup>, 5.80 kJ mol<sup>-1</sup>, 97.4 J mol<sup>-1</sup> K<sup>-1</sup> for humic acid, and from -19.6 kJ mol<sup>-1</sup>, 4.43 kJ mol<sup>-1</sup>, 80.1 J mol<sup>-1</sup> K<sup>-1</sup> to -25.1 kJ mol<sup>-1</sup>, 7.50 kJ

mol<sup>-1</sup>, 96.4 J mol<sup>-1</sup> K<sup>-1</sup> for humin, respectively. The negative  $\Delta G^\circ$  values at all temperatures indicated the feasibility and spontaneous nature of the adsorption of Cu<sup>2+</sup> on humic acid and humin, which means that the adsorptive force is strong enough to break the potential and lead the reaction to bind Cu<sup>2+</sup> onto the surface functional groups of humic substance fractions (Li *et al.*, 2008). The decrease in  $\Delta G^\circ$  values with increase in temperature implied that the adsorption process was more favorable at higher temperature. The positive  $\Delta H^\circ$  values indicated that the adsorption reaction was an endothermic process, which is supported by the results that the adsorption of Cu<sup>2+</sup> on humic acid and humin increased with increasing temperature. The positive  $\Delta S^\circ$  values reflected the affinity of the humic substance fractions towards Cu<sup>2+</sup> and the increase of the degree of freedom at the solid/liquid interface during the adsorption process. In the study of Prado and Airoidi (2003), they showed that the adsorption of Cu<sup>2+</sup> on

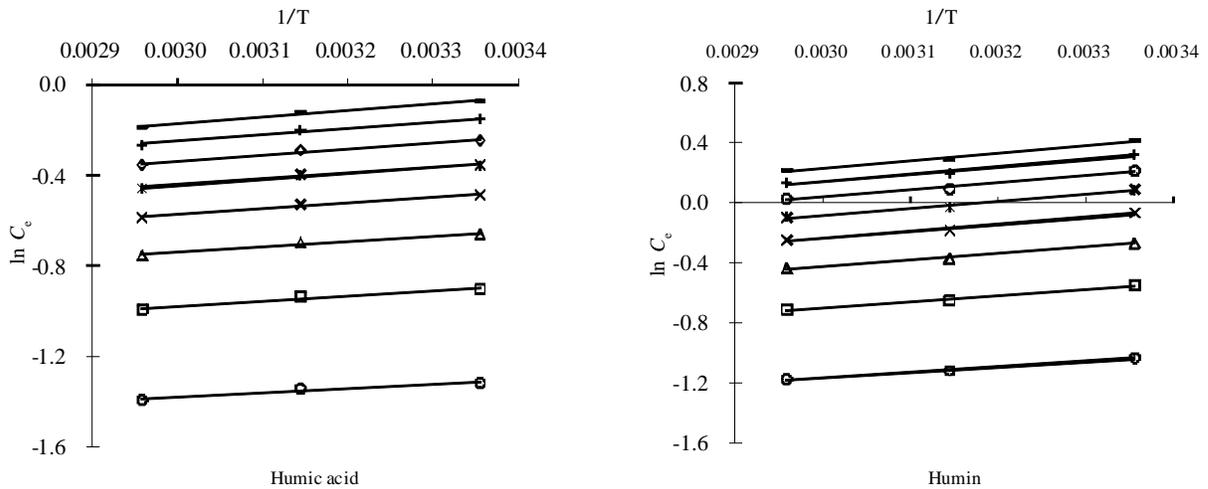


Fig. 4. Linear plot of  $\ln C_e$  vs.  $1/T$  for different amounts of  $\text{Cu}^{2+}$  adsorption on humic acid and humin,  $q_e=2$  ( $\circ$ ), 4 ( $\square$ ), 6 ( $\triangle$ ), 8 ( $\times$ ), 10 ( $*$ ), 12 ( $\diamond$ ), 14 ( $+$ ), 16 ( $-$ )  $\text{mg g}^{-1}$ , from black soil

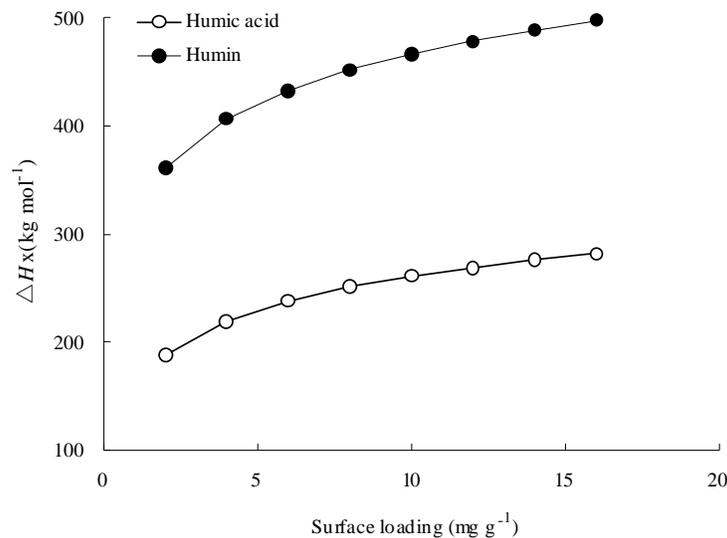


Fig. 5. Plot of isosteric heat of adsorption ( $\Delta H_x$ ) against surface loading for the adsorption of  $\text{Cu}^{2+}$  on humic acid and humin from black soil

humic acids from peat soil and commercial Aldrich was spontaneous, endothermic and increasingly disordered process, in accordance with our results.

Compared with humic acid, the  $\Delta G^\circ$  values of  $\text{Cu}^{2+}$  adsorption on humin were always more positive at each  $\text{Cu}^{2+}$  concentrations studied. It indicated that the adsorption of  $\text{Cu}^{2+}$  was more favorable on humic acid, in agreement with the larger values of the Freundlich constant  $n$  of humic acid. However, there was no consistent difference of the values of  $\Delta H^\circ$  and  $\Delta S^\circ$  between humic acid and humin.

The plots of  $\ln C_e$  versus  $1/T$  are plotted in Fig. 4, and the corresponding  $\Delta H_x$  values are listed in Table

6. The plots of  $\ln C_e$  versus  $1/T$  were found to be linear with  $R^2$  values greater than 0.943. The  $\Delta H_x$  values ranged from 188.1 to 281.7  $\text{kJ mol}^{-1}$  for humic acid and from 361.0 to 497.4  $\text{kJ mol}^{-1}$ . The values of  $\Delta H_x$  for physical and chemical adsorption are below 80  $\text{kJ mol}^{-1}$  and between 80 and 400  $\text{kJ mol}^{-1}$ , respectively (Chowdhurg *et al.*, 2011). Thus, our present results implied that the chemical adsorption involving ion exchange was the dominating mechanism. On the other hand, the variation of  $\Delta H_x$  with surface loading (Fig. 5) showed that the  $\Delta H_x$  values increased with the increase of  $q_e$  values, indicating that the adsorbents had energetically heterogeneous surfaces. The above variation in  $\Delta H_x$

**Table 6. Isothermic heat of adsorption ( $\Delta H_x$ ) of Cu<sup>2+</sup> on humic acid and humin from black soil**

Samples	$q_e$ (mg g <sup>-1</sup> )	$\Delta H_x$ (kJ mol <sup>-1</sup> )	$R^2$
Humic acid			
	2	188.1	0.943
	4	219.3	0.964
	6	237.5	0.972
	8	250.5	0.977
	10	260.5	0.980
	12	268.8	0.982
	14	275.7	0.984
	16	281.7	0.985
Humin			
	2	361.0	0.993
	4	406.5	0.989
	6	433.1	0.987
	8	451.9	0.985
	10	466.6	0.984
	12	478.5	0.983
	14	488.6	0.982
	16	497.4	0.982

values was usually due to the adsorbate-adsorbate interaction followed by the adsorbate-adsorbent interaction, and the possibility of having lateral interactions between adsorbed Cu<sup>2+</sup> ions (Chowdhurg *et al.*, 2011; Srivastava *et al.*, 2007). The extent of the increase in  $\Delta H_x$  values was larger for humic acid (49.8%) than for humin (37.8%), implying that humic acid had greater surface energy heterogeneity than humin. It agreed with the larger  $n$  values of humic acid with respect to humin in the Freundlich constant. In the study of Li *et al.* (2004), they indicated that the humic substance fractions with lower molecular weights, greater O/C atomic ratios, and higher contents of oxygen and lignin-derived aromatic structural units were more heterogeneous. Based on our present study, the O/C atomic ratio and the contents of oxygen and lignin-derived aromatic structural units of humic acid were all higher than that of humin. Meanwhile, the molecular weight of humic acid was generally considered lower than that of humin (Stevenson, 1994). It confirmed that the heterogeneity of humic acid was larger than that of humin. Thus, the difference in heterogeneity of adsorption surface of humic acid and humin could be due on their different chemical composition.

## CONCLUSIONS

The chemical compositions of humin differed from those of corresponding humic acid. The former was more humified, aliphatic, hydrophobic, and less polar compared with the latter. The adsorption amounts of humin for Cu<sup>2+</sup> were lower than that of humic acid.

The maximum adsorption capacities of Cu<sup>2+</sup> on humin and humic acid were 227.3 and 243.9 mg g<sup>-1</sup> at 298 K, respectively. The adsorption data could be better described Freundlich equation with respect to equation. The increase in temperature had a positive effect on the sorption process. The adsorption of Cu<sup>2+</sup> on humin and humic acid was a spontaneous, endothermic, and increasingly disordered process. Both humin and humic acid had energetically heterogeneous surfaces for the adsorption of Cu<sup>2+</sup>, and the surface energy heterogeneity of humic acid was greater than that of humin.

## ACKNOWLEDGEMENTS

This work was financially supported by the National Key Technology R&D Program (2013BAD07B02 and 2013BAC09B01), the National Agricultural Department Public Benefit Research Foundation (201103003), and the Postdoctoral Project of Jilin Province (01912). The authors are grateful to Dr. Zijiang Jiang of the Changchun Institute of Applied Chemistry, Chinese Academy of Sciences, for his technical assistance in obtaining the solid-state <sup>13</sup>C NMR spectra.

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