

## Internal Loads of Nutrients in Lake Chaohu of China: Implications for lake Eutrophication

Yang, L.<sup>1,2</sup>, Lei, K.<sup>1</sup>, Yan, W.<sup>2</sup> and Li, Y.<sup>3\*</sup>

<sup>1</sup>River and Coastal Environment Research Center, Chinese Research Academy of Environmental Sciences, Beijing 100012, P. R. China

<sup>2</sup>Institute of Geographic Sciences and Natural Resources Research, Chinese Academy of Sciences, 11A Datun Road, Beijing, 100101, P.R. China

<sup>3</sup>Anhui University, Hefei, 230039, P.R. China

Received 2 Nov. 2012;

Revised 10 June 2013;

Accepted 17 June 2013

**ABSTRACT:** This study investigated the impact of internal nitrogen and phosphorus loading on water quality of Lake Chaohu, China. Results showed that the overall mean concentration of TN and TP in lake water was  $1.09 \pm 0.84$  and  $0.15 \pm 0.84$  (SD) mg/L, respectively, indicating Lake Chaohu was highly eutrophic during the study period. Mean benthic flux of  $\text{NH}_4^+$ ,  $\text{NO}_3^-$ , and  $\text{PO}_4^{3-}$  were  $3.56 \pm 1.52$  mg N/m<sup>2</sup>/d,  $-0.14 \pm 0.40$  mg N/m<sup>2</sup>/d, and  $-1.54 \pm 21.94$   $\mu\text{g}$  P/m<sup>2</sup>/d in summer and were  $3.16 \pm 1.83$  mg N/m<sup>2</sup>/d,  $-0.03 \pm 0.30$  mg N/m<sup>2</sup>/d, and  $-0.11 \pm 2.58$   $\mu\text{g}$  P/m<sup>2</sup>/d in fall, respectively. In this study, lake sediments acted as both a source of  $\text{NH}_4^+$  and a sink of  $\text{PO}_4^{3-}$ . A scenario was used to understand the role of sediments on lake water quality. Results showed that Lake Chaohu would be eutrophic within a month under the impact of internal  $\text{NH}_4^+$  loading when water was diverted to the lake. This indicates that an integrated project combining both a diversion project and dredging of sediments would be necessary for future lake restoration.

**Key words:** Benthic fluxes, Lake Chaohu, Nitrogen, Phosphorus, Eutrophication

### INTRODUCTION

Lake eutrophication caused by excessive inputs of phosphorus and nitrogen is a worldwide problem, and can lead to water quality deterioration with significant losses of biodiversity, goods and services (Akdeniz *et al.*, 2011; Raicevic *et al.*, 2011; Schindler 2006; Dokulil and Teubner 2006). Lake sediments can have higher nitrogen and phosphorus levels due to the decomposition of aquatic flora and fauna (Gerhardt *et al.*, 2010; Graca *et al.*, 2004). Nutrients in lake sediments can be released into the water column due to physical, chemical and biological processes, which is termed internal loading, contrasting to external loading (Pettersen 1998). As a result, lake sediments may play an important role in nutrient cycling and water eutrophication (Håkanson 2004). For instance, external nitrogen and phosphorus loads have been reduced by 40-80% in China during recent decades (Xu *et al.*, 2003); however, more than 90% of Chinese lakes are still eutrophic. Thus, an evaluation of the impact of internal loading on water quality is necessary when a lake recovery strategy is considered. Lake Chaohu in China

once played an important role in supporting local development (Yang *et al.*, 2011). However, this lake has been eutrophic since the 1970s due to increased nutrient inputs (Shang and Shang 2007). Since 2006, the Chinese government implemented a diversion project, called "Transferring Water from the Yangtze River to Lake Chaohu (TWFYTC)," as an emergency strategy to control the lake eutrophication. Unfortunately, water quality in Lake Chaohu showed no significant improvement. Thus, research, monitoring, and evaluation activities are essential for characterizing eutrophication in this lake. There is especially a need to understand the role of internal nutrient loading in lake water quality. To the best of our knowledge, previous studies report on nitrogen and phosphorus distribution in water and sediment profiles in Lake Chaohu (Pan *et al.*, 2007), while *in situ* measurements on benthic fluxes of nutrients across the sediment-water interface are sparse. In this study, *in situ* measurements were conducted in June and September 2008 to investigate benthic fluxes of nitrogen and phosphorus in Lake Chaohu. In addition,

\*Corresponding author E-mail: lbyang123@126.com

we designed a scenario to model the impact of internal nitrogen and phosphorus loading on lake water quality. This study will provide valuable information that can be used for control of eutrophication in the future.

## MATERIALS & METHODS

Twelve sampling sites were selected throughout Lake Chaohu (Fig. 1). Sediment (5 cm) samples were collected with a Piston sampler and water samples were collected with a Niskin water sampler. The upper-layer sediment samples were centrifuged at 4500 r/m to obtain the interstitial water of the sediments. The centrifuged water samples were filtrated through a 0.45 $\mu$ m filter and the concentration of  $\text{NH}_4^+$ ,  $\text{NO}_3^-$ ,  $\text{PO}_4^{3-}$ , TN, and TP were measured with a segmented flow auto-analyzer. In the non-centrifuged water samples, TN and TP were determined using a digestion method (Mckee *et al.*, 2000; Venterink *et al.*, 2003). TN, TP, and organic matter in the lake sediment samples were measured according to Chen *et al.* (2010).

Benthic fluxes of  $\text{NH}_4^+$ ,  $\text{NO}_3^-$  and  $\text{PO}_4^{3-}$  were calculated using the equation given by Berner (1980):

$$F = -D_s \Phi (\partial C / \partial x)_{x=0} \quad (1)$$

where F is the benthic flux of nutrients ( $\text{mmol}/\text{m}^2/\text{h}$ );  $D_s$  is the molecular diffusion coefficient ( $\text{cm}^2/\text{s}$ ) calculated as in Krom & Berner (1980);  $\Phi$  is the sediment porosity (%) calculated by the equation  $\Phi = W / \{(100 - W)\rho + W\}$ , W and  $\rho$  are the sediment moisture content (%) and the bulk density ( $\text{g}/\text{cm}^3$ ), respectively;  $(\partial C / \partial x)_{x=0}$  ( $\text{mmol}/\text{L}/\text{m}$ ) is the concentration gradient of nutrients between the sediment's interstitial and overlying water. Here, a negative flux means that the water nutrients were

consumed by sediments and a positive flux means that the nutrients were released from the sediments to the water column.

## RESULTS & DISCUSSION

Our data showed that Lake Chaohu was eutrophic during the experimental period. Overall, concentrations of  $\text{NO}_3^-$ ,  $\text{NH}_4^+$ ,  $\text{PO}_4^{3-}$ , TN, and TP in surface water ranged from 0.08~1.93, 0.05~1.81, 0.02~0.16, 0.25~3.67, and 0.07~0.29 mg/L (Table 1), respectively. Nitrogen and phosphorus concentrations showed no significant seasonal differences. Significantly higher levels of nitrogen and phosphorus in the western lake indicated spatial variation in Lake Chaohu. This is a result of large amounts of nitrogen and phosphorus discharge into the western lake by two major badly polluted tributaries, the Nanfei and Hangbu Rivers (Shang and Shang 2007).

Three sediment types (clay, sand, and mud) were investigated on a whole lake basis (Table 2). Data showed that lake sediments were enriched with organic matter, TN, and TP, with a range of 0.56~1.51%, 0.22~0.92 g/kg, and 0.11~0.66 g/kg, respectively. Porosity of lake sediments exhibited a range of 11.8~79.3%, with higher porosity in sandy sediments. No differences in seasonal or spatial distribution of sediment characters were found in this lake.

The benthic flux of nitrogen and phosphorus was calculated according to equation 1 based on the concentration gradient of nutrients (3% concentration) between interstitial and overlying water of sediments. Concentrations of  $\text{NH}_4^+$ ,  $\text{NO}_3^-$ , and  $\text{PO}_4^{3-}$  in the sediment interstitial and overlying water are presented in Fig.2.

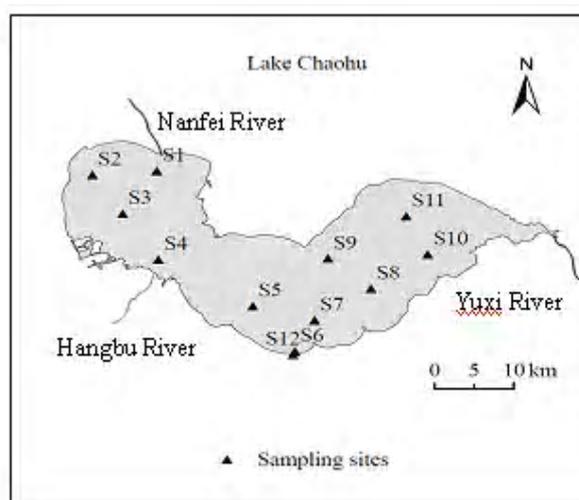


Fig. 1. Location of sampling sites in Lake Chaohu

**Table 1. Nitrogen and phosphorus concentrations in surface lake water (mg/L)**

Site	NO <sub>3</sub> <sup>-</sup>		NH <sub>4</sub> <sup>+</sup>		PO <sub>4</sub> <sup>3-</sup>		TN		TP	
	Jun.	Sep.	Jun.	Sep.	Jun.	Sep.	Jun.	Sep.	Jun.	Sep.
S1	0.65	1.93	0.46	1.81	0.16	0.09	1.67	3.67	0.29	0.24
S2	0.45	0.94	0.05	0.22	0.02	0.07	2.12	1.62	0.20	0.26
S3	0.77	0.35	0.09	0.23	0.03	0.05	3.27	1.05	0.24	0.21
S4	0.47	0.22	0.19	0.31	0.02	0.05	0.92	0.69	0.14	0.07
S5	0.10	0.13	0.16	0.12	0.03	0.04	0.83	0.80	0.14	0.07
S6	0.08	0.17	0.16	0.13	0.03	0.03	0.66	0.57	0.18	0.09
S7	0.08	0.11	0.23	0.10	0.02	0.04	0.84	1.00	0.12	0.16
S8	0.09	0.11	0.25	0.16	0.02	0.04	0.70	0.25	0.11	0.15
S9	0.08	0.12	0.49	0.16	0.02	0.04	0.58	0.70	0.11	0.18
S10	0.08	0.11	0.08	0.19	0.05	0.03	0.63	0.45	0.10	0.14
S11	0.08	0.08	0.07	0.21	0.02	0.05	0.82	0.63	0.09	0.11
S12	0.18	0.27	0.12	0.13	0.02	0.03	0.97	0.68	0.12	0.19
Mean	0.26±0.25	0.38±0.54	0.20±0.14	0.31±0.47	0.03±0.04	0.05±0.02	1.17±0.81	1.01±0.91	0.15±0.06	0.16±0.06

Data are presented as “mean±SD”

**Table 2. Sediment characterization data of Lake Chaohu**

site	Sediments type	Porosity (%)		Organic matter (%)		TN (g/kg)		TP (g/kg)	
		Jun.	Sep.	Jun.	Sep.	Jun.	Sep.	Jun.	Sep.
S1	clay, mud	64.3		1.51		0.54		0.37	
S2	clay, mud	62.9		1.13		0.63		0.49	
S3	sand	70.9		0.71		0.73		0.61	
S4	clay, mud	49.8		0.91		0.92		0.66	
S5	sand	79.3	75.6	0.64	0.71	0.81	0.78	0.63	0.65
S6	clay	37.5		1.32		0.86		0.64	
S7	mud	42.2	40.8	0.91	1.21	0.22	0.92	0.14	0.15
S8	sand, mud	73.5	72.5	0.65	0.55	0.39	0.76	0.11	0.13
S9	sand, mud	70.2	68.6	0.62	0.72	0.31	0.53	0.13	0.11
S10	mud	52.8	50.3	0.87	0.91	0.35	0.54	0.20	0.16
S11	sand, mud	78.1	72.6	0.56	0.76	0.39	0.87	0.27	0.12
S12	clay	11.8		1.17		0.37		0.24	
Mean of all samples		57.8±19.9	63.4±14.3	0.92±0.31	0.81±0.23	0.54±0.24	0.73±0.16	0.37±0.22	0.22±0.21

Data are presented as “mean±SD”, six sampling sites (S5, S7-S11) were sampled in September.

Overall, NH<sub>4</sub><sup>+</sup> concentrations in the sediment’s interstitial water were higher than the overlying water, while NO<sub>3</sub><sup>-</sup> and PO<sub>4</sub><sup>3-</sup> concentrations in the sediment’s interstitial water were lower than that in the overlying water at most of the sampling sites.

Benthic fluxes of NH<sub>4</sub><sup>+</sup>, NO<sub>3</sub><sup>-</sup>, and PO<sub>4</sub><sup>3-</sup> ranged from 0.19~8.26 mg N/m<sup>2</sup>/d (mean 3.56), -0.83~0.05 mg N/m<sup>2</sup>/d (mean -0.14), and -27.46~6.27 μg P/m<sup>2</sup>/d (mean -1.54) in June (Fig. 3), respectively. Six sampling sites (S5, S7-S11) were sampled in September. Benthic fluxes of NH<sub>4</sub><sup>+</sup>, NO<sub>3</sub><sup>-</sup> and PO<sub>4</sub><sup>3-</sup> in September ranged from 0.87~5.84 mg N/m<sup>2</sup>/d, -0.18~0.03 mg N/m<sup>2</sup>/d, and -10.61~ -3.77 μg P/m<sup>2</sup>/d, with mean values of 3.16 mg N/m<sup>2</sup>/d, -

0.04 mg N/m<sup>2</sup>/d, and -6.19 μg P/m<sup>2</sup>/d, respectively (Table 3). The overall mean PO<sub>4</sub><sup>3-</sup> flux of the six sampling sites in September was significantly higher than that in June (One-way ANOVA; α = 0.05, P = 0.002).

Compared to other studies, benthic fluxes of nitrogen and phosphorus in Lake Chaohu were relatively lower. For example, Höhener and Gächter (1994) reported that the benthic flux of NH<sub>4</sub><sup>+</sup> ranged from 15.4~225.4 mg N/m<sup>2</sup>/h, and the benthic flux of NO<sub>3</sub><sup>-</sup> ranged from -33.6~155.4 mg N/m<sup>2</sup>/h in Lake Sempach. Qin and Zhu (2006) surveyed the lakes from the middle to lower reaches of the Yangtze River basin, and found that mean TN and NH<sub>4</sub><sup>+</sup> release rates were 166 and 46

Internal loads of nutrients in Lake Chaohu

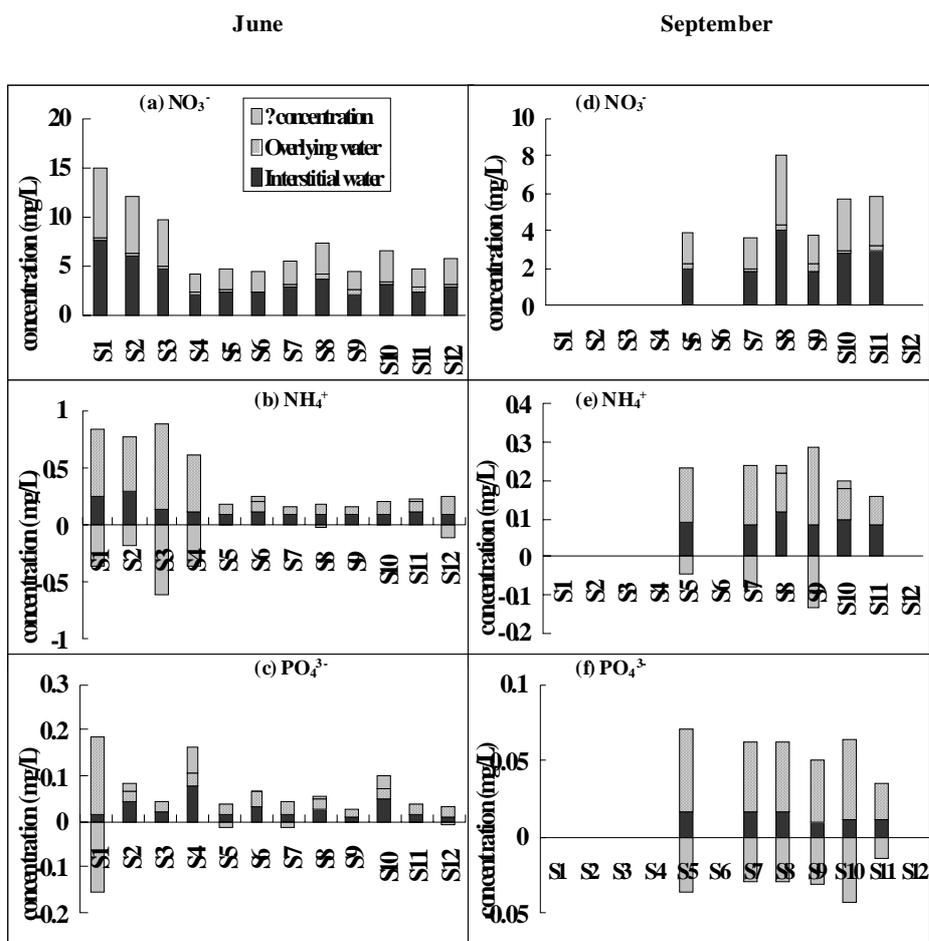


Fig. 2. Concentration of  $\text{NO}_3^-$ ,  $\text{NH}_4^+$  and  $\text{PO}_4^{3-}$  in interstitial and overlying water in Lake Chaohu

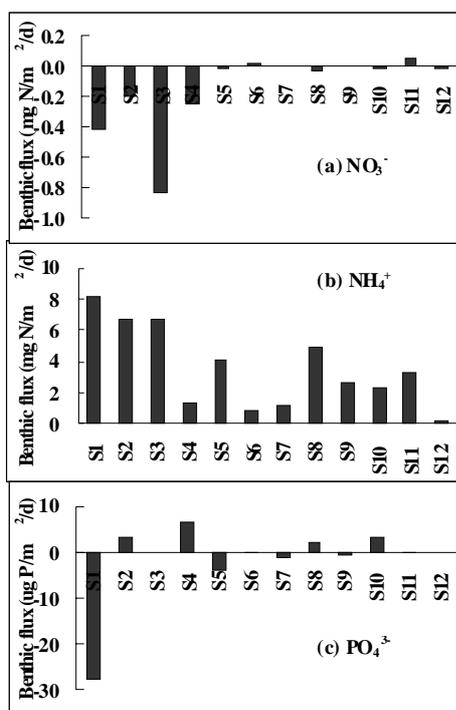


Fig. 3. Benthic flux of  $\text{NO}_3^-$ ,  $\text{NH}_4^+$  and  $\text{PO}_4^{3-}$  in Lake Chaohu in June

**Table 3. Benthic flux of nitrogen and phosphorus in Lake Chaohu in September**

Observation time	Sampling site	Benthic flux		
		NH <sub>4</sub> <sup>+</sup> mg N/m <sup>2</sup> /d	NO <sub>3</sub> <sup>-</sup>	PO <sub>4</sub> <sup>3-</sup> µg P/m <sup>2</sup> /d
September	S5	3.24	-0.09	-10.61
	S7	0.87	0.01	-3.77
	S8	5.84	0.03	-6.98
	S9	2.19	-0.18	-6.74
	S10	2.11	0.02	-5.15
	S11	4.68	0.01	-3.87
	Mean	3.16±1.83	-0.03±0.08	-6.19±2.56
June	Mean	3.09±1.30	-0.03±0.30	-0.11±2.58
Significance test	P	0.95	1.00	0.002*

Mean values are presented as “Mean ± SD”; \*indicates a significant difference at 0.05 test level

mg N/m<sup>2</sup>/h, respectively. In other reports, PO<sub>4</sub><sup>3-</sup> fluxes ranged from -53~937 mg P/m<sup>2</sup>/d (Gibbs and Özkundakci 2010; Perrone *et al.*, 2008; Punning and Kapanen, 2009; Penn *et al.*, 2000), and our results occur at the bottom of this range.

Results indicate that sediments in Lake Chaohu act both as a source of nitrogen and as a sink of phosphorus. This is consistent with other reports on various lakes. When taking the total area of lake sediment (750 km<sup>2</sup>) into consideration, it is estimated that approximately 925 tons of NH<sub>4</sub><sup>+</sup> were released to the water column from sediments while 0.87 tons of PO<sub>4</sub><sup>3-</sup> were consumed by sediment annually. Existing data had shown that approximately 9500 tons of TN and 490 tons of TP were discharged to the Lake Chaohu annually. Thus, about 9.7% of the nitrogen input to the lake was contributed by internal NH<sub>4</sub><sup>+</sup> loading while 0.2% of the phosphorus was consumed by sediments in the form of PO<sub>4</sub><sup>3-</sup>.

It has become evident that nutrient exchange between sediments and water is highly complex and includes interrelated chemical, biological and physical processes (Holdren and Armstrong, 1980). A variety of ecological responses in lakes would result in nutrient exchanges from the sediments to the overlying water column. Consequently, a number of environmental factors are important in benthic nutrients fluxes in aquatic ecosystems (Pettersson, 1998; Malecki *et al.*, 2004). In our study, benthic flux of NH<sub>4</sub><sup>+</sup> was significantly correlated with porosity of sediment (R<sup>2</sup>=0.38, P=0.006; Table 4), but there were no significant correlations between benthic nutrient fluxes and other measured environmental factors. This may indicate that other environmental factors, such as

sediment bacteria and aquatic plankton, could regulate the exchange of nitrogen and phosphorus in Lake Chaohu. This has been reported in other lakes. For example, Pettersson (1998) found that the cyanobacteria *Gloeotrichia echinulata* contributed significantly to internal loading of phosphorus in Lake Erken. In Lake Vallentunasjön, Boström *et al* (1988) reported that microbial metabolism had a number of direct effects on phosphorus mobility, which would control the exchange of phosphorus between the sediments and the overlying water column. Thus, it is necessary that more detailed biogeochemical data be obtained in future studies in order to distinguish the dominant regulatory factors on benthic nutrient fluxes in Lake Chaohu.

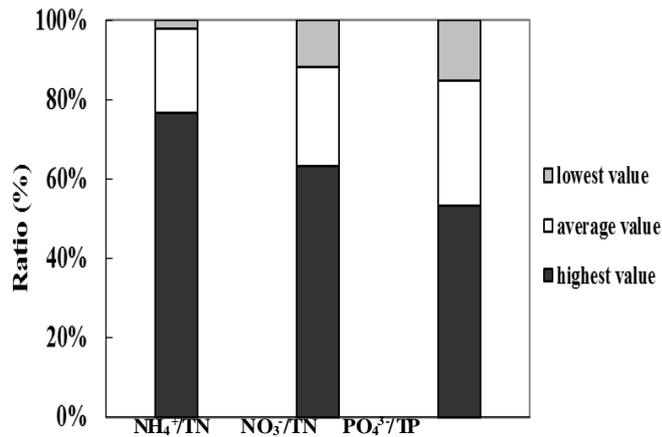
Many studies have supported the hypothesis that internal loading can often determine the eutrophication status of a lake and that there is a time lag for recovery after the reduction of external loading (Søndergaard *et al.*, 2003; Mitamura *et al.*, 1995). Thus, even if only a very small amount is released, it will have a significant impact on lake water quality (Présing *et al.*, 2001). In this article, a scenario was designed to model the potential risk of eutrophication by internal loading. The “TWFYTC” project was performed twice a year from 2006 to the present, and was sustained for several days each time. Therefore, our scenario was made based on the assumptions that: 1) lake water was completely replaced by the transferred water; 2) external loading had no significant impact on water quality during the “TWFYTC” project. Here, the implication of internal loading for lake water quality was studied by calculating the recurrence time (T) of eutrophication:

$$T = (V \times C) / (A \times F) \quad (2)$$

**Table 4. Correlation analysis between benthic nutrients fluxes and environmental factors**

Dependent	Independent	R <sup>2</sup>	P
NH <sub>4</sub> <sup>+</sup> flux	Water temperature	0.06	0.33
	TN content in sediment	0.003	0.82
	Organic matter in sediment	0.01	0.77
	Porosity of sediment	0.38	0.006*
NO <sub>3</sub> <sup>-</sup> flux	Water temperature	0.14	0.12
	TN content in sediment	0.01	0.66
	Organic matter in sediment	0.01	0.78
	Porosity of sediment	0.02	0.59
PO <sub>4</sub> <sup>3-</sup> flux	Water temperature	0.02	0.59
	TP content in sediment	0.003	0.86
	Organic matter in sediment	0.10	0.19
	Porosity of sediment	0.05	0.38

\*indicates a significant correlation at 0.05 test level (n=18)



**Fig. 4. Ratios of NH<sub>4</sub><sup>+</sup>/TN, NO<sub>3</sub><sup>-</sup>/TN, and PO<sub>4</sub><sup>3-</sup>/TP in lake water**

**Table 5 . Assessment of eutrophication risk from by internal loading in Lake Chaohu**

	C <sub>S</sub> (mg/L)	C <sub>W</sub> (mg/L)	F (mg/m <sup>2</sup> /d)	CL (ton)	T
NH <sub>4</sub> <sup>+</sup>		0.045	3.43	1.7	31 (d)
	3.46	0.030	3.44	27.5	31 (d)
		0.015	3.46	53.3	30 (d)
NO <sub>3</sub> <sup>-</sup>		0.036	0.09	6.9	2.8 (yr)
	0.13	0.024	0.10	27.5	2.5 (yr)
		0.012	0.11	48.2	2.3 (yr)
PO <sub>4</sub> <sup>3-</sup>		0.009	0.006	3.4	7.8 (yr)
	0.03	0.006	0.007	8.6	7.4 (yr)
		0.003	0.008	13.8	7.2 (yr)

Where  $V$  is the lake storage capacity ( $1.72 \times 10^9 \text{ m}^3$ );  $A$  is the lake sediment area ( $750 \text{ km}^2$ );  $F$  is the benthic flux of nutrients;  $C$  is the critical limit concentration of TN ( $0.2 \text{ mg/L}$ ) and TP ( $0.02 \text{ mg/L}$ ) meets the definition of eutrophication in a lake (according to Chinese standard methods for a lake eutrophication survey) (Jin and Tu 1990). In this study, the ratios of  $\text{NH}_4^+/\text{TN}$ ,  $\text{NO}_3^-/\text{TN}$  and  $\text{PO}_4^{3-}/\text{TP}$  in lake water were 23%, 20% and 57% (Fig. 4), respectively. Therefore, we assumed that Lake Chaohu would be eutrophic when  $\text{NH}_4^+$ ,  $\text{NO}_3^-$  and  $\text{PO}_4^{3-}$  had reached the thresholds of  $0.046$ ,  $0.04$  and  $0.011 \text{ mg/L}$ , respectively.

Three concentration levels for  $\text{NH}_4^+$ ,  $\text{NO}_3^-$  and  $\text{PO}_4^{3-}$  in the transferred water were modeled. Results showed that when concentrations of  $\text{NH}_4^+$ ,  $\text{NO}_3^-$  and  $\text{PO}_4^{3-}$  in the transferred water varied from  $0.015$  to  $0.045$ ,  $0.012$  to  $0.036$ ,  $0.003$  to  $0.009 \text{ mg/L}$ , respectively, the lake water would be eutrophic within about a month (Table 5). We can also conclude that internal  $\text{NH}_4^+$  loading regulates the eutrophication status of Lake Chaohu. However, it is important to note that the impact of internal loading on the lake eutrophic condition might be overestimated since the reduction processes of the nutrients were not considered. Monitoring data showed that monthly TN and TP concentrations ranged from  $0.54$ – $6.59$  and  $0.08$ – $0.56 \text{ mg/L}$ , respectively, from 2006 to the present in Lake Chaohu (unpublished data), meaning this lake was still eutrophic. This is consistent with the result of our scenario that Lake Chaohu would be eutrophic within a month after the “TWFYTC” project would be performed. This strongly suggests that an integrated project combining both a diversion project and dredging of sediments would be necessary for improving the water quality of Lake Chaohu in the future.

$C_s$  is the mean concentration of nitrogen and phosphorus in sediment interstitial water from this study;  $F$  is the benthic flux of nitrogen and phosphorus calculated according to equation 1.  $CL$  is the critical limit internal loading of nitrogen and phosphorus that meets the definition of lake eutrophication.

## CONCLUSION

The overall mean concentration of TN and TP in the lake water was  $1.09 \pm 0.84$  and  $0.15 \pm 0.84$  (SD)  $\text{mg/L}$ , respectively, indicating that Lake Chaohu was highly eutrophic. Lake sediments acted as both a source of  $\text{NH}_4^+$  and a sink of phosphate. It is estimated that about 9.7% of the nitrogen input to the lake was contributed by internal  $\text{NH}_4^+$  loading while 0.2% of phosphorus was consumed by sediments in the form of  $\text{PO}_4^{3-}$ . A scenario showed that Lake Chaohu tends to be eutrophic within about a month under the influence of internal  $\text{NH}_4^+$  loading after the “TWFYTC” project would be

performed. This strongly suggests that an integrated project combining both a diversion project and dredging of sediments would be necessary for the control of eutrophication.

## ACKNOWLEDGEMENTS

This work was supported by the National Science-technology Support Project of China (No. 2010BAC69B02) and National Natural Science Foundation of China (No.41203080).

## REFERENCES

- Akdeniz, S., Karaer, F., Katip, A. and Aksoy, E. (2011). A Gis-based method for shallow lake eutrophication assessment. *J. Biol. Environ. Sci.*, **5**(15), 195-202.
- Berner, R. A. (1980). Early Diagnosis: A Theoretical Approach. In Berner, R. A. eds, Princeton, New Jersey.
- Boström, B., Andersen, J. M., Fleischer, S. and Jansson, M. (1988). Exchange of Phosphorus across the Sediment-Water Interface. *Hydrobiologia*, **170** (1), 229-244.
- Chen, R. H., Zhan, L. T., Chen, Y. M. and Hu, H. Z. (2010). Contents of nitrogen, phosphorus and organic materials in sediments and their distribution along depth at Xixi wetland. *China Environmental Science*, **30** (4), 493-498. (In Chinese)
- Dokulil, M. T. and Teubner K. (2003). Eutrophication and restoration of shallow lakes: is the concept of alternative stable equilibria a useful tool. *Hydrobiologia*, **506-509** (1-3), 29-35.
- Gibbs, M. and Özkundakci, D. (2010). Effects of a modified zeolite on p and n processes and fluxes across the lake sediment-water interface using core incubations. *Hydrobiologia*, **661** (1), 21-35.
- Graca, B., Burska, D. and Matuszewska, K. (2004). The impact of dredging deep pits on organic matter decomposition in sediments. *Water, Air and Soil Pollution*, **158** (1), 237-259.
- Gerhardt, S., Boos, K. and Schink, B. (2010). Uptake and release of phosphate by littoral sediment of a freshwater lake under the influence of light or mechanical perturbation. *Journal of Limnology*, **69** (1), 54-63.
- Håkanson, L. (2004). Internal loading: a new solution to an old problem in aquatic sciences. *Lakes & Reservoirs*, **19**, 3-23.
- Höhener, P. and Gächter, R. (1994). Nitrogen Cycling across the Sediment-Water Interface in An Eutrophic, Artificially Oxygenated Lake. *Aquatic Sciences*, **56**, 115-132.
- Holdren, G. C. and Armstrong, D. E. (1980). Factors affecting phosphorus release from intact lake sediment cores. *Environ. Sci. Technol.*, **14** (1), 79-87.
- Jin, X. C. and Tu, Q. Y. (1990). *The Standard Methods for Observation and Analysis in Lake Eutrophication*, 2nd ed. Beijing.

- Krom, M. D. and Berner, R. A. (1980) Adsorption of phosphate in anoxic marine sediments. *Limnology and Oceanography*, **25** (5), 797-806.
- Mckee, L. J., Eyre, B. D. and Hossain, S. (2000). Transport and retention of nitrogen and phosphorus in the sub-tropical richmond river estuary, australia-a budget approach. *Biogeochemistry*, **50**, 241-278.
- Mitamura, O., Saijo, Y., Hino, K. and Barbosa, F. A. R. (1995). The Significance of regenerated Nitrogen for Phytoplankton productivity in the Rio Doce Valley Lakes, Brazil. *Archiv für Hydrobiologie*, **134** (4), 179-194.
- Malecki, L. M., White, J. R. and Reddy, K. R. (2004). Nitrogen and phosphorus flux rates from sediment in the Lower St. Johns River Estuary. *J. Environ. Qual.*, **33** (4), 1545-1555.
- Petterson, K. (1998). Mechanisms for internal loading of phosphorus in lakes. *Hydrobiologia* **373/374**, 21-25.
- Perrone, U., Facchinelli, A. and Sacchi, E. (2008). Phosphorus dynamics in a small eutrophic Italian lake. *Water, Air, & Soil Pollution.*, **189** (1-4), 335-351.
- Penn, M. R., Auer, M. T., Doerr, S. M. and Driscoll, C. T. (2000). Seasonality in Phosphorus Release Rates from the Sediments of a Hypereutrophic Lake under a Matrix of pH and Redox Conditions. *Canada journal of fisheries and aquatic science*, **57** (5), 1033-1041.
- Présing, M., Herodek, S., Preston, T. and Voros, L. (2001). Nitrogen uptake and the importance of internal nitrogen loading in Lake Balaton. *Freshwater Biology*, **46** (1), 125-139.
- Punning, J. M. and Kapanen, G. (2009). Phosphorus flux in Lake Peipsi sensu stricto, Eastern Europe. *Estonian Journal of Ecology*, **58**, 1, 3-17.
- Pan, C. R., Wang, J. Q., Zheng, Z. X., Liu, J. J. and Yin, F. C. (2007). Forms of phosphorus and nitrogen existing in sediments in Chaohu Lake. *Journal of Ecology and Rural Environment*, **23** (1), 43-47. (In Chinese)
- Qin, B. Q. and Zhu, G. W. (2006). The nutrients forms, cycling and exchange flux in the sediment and overlying water system in lakes from the middle and lower reaches of Yangtze River. *Science in China*, **49** (Supp.1), 1-13.
- Raicevic, V., Bozic, M., Rudic, Z., Lalevic, B. and Kikovic, D. (2011). The evolution of the eutrophication of the Palić Lake (Serbia). *African Journal of Biotechnology*, **10** (10), 1736-1744.
- Søndergaard, M., Jensen, J. P. and Jeppesen, E. (2003). Role of sediment and internal loading of phosphorus in shallow lakes. *Hydrobiologia*, **506-509** (1-3), 135-145.
- Shang, G. P. and Shang, J. C. (2007). Spatial and Temporal Variations of Eutrophication in Western Chaohu Lake, China. *Environmental Monitoring and Assessment*, **130** (1-3), 99-109.
- Schindler, D.W. (2006). Recent advances in the understanding and management of eutrophication. *Limnol Oceanogr.*, **51** (1, part 2), 356-363.
- Venterink, H.O., Wiegman, F., Van der Lee, G. E. M. and Vermaat, J. E. (2003). Role of active floodplains for nutrients retention in the river Rhine. *Journal of Environmental Quality*, **32** (4), 1430-1435.
- Xu, F. L., Tao, S., Dawson, R. W., and Xu, Z. R. (2003). The distributions and effects of nutrients in the sediments of a shallow eutrophic Chinese lake. *Hydrobiologia*, **429** (1-3), 85-93.
- Yang, L. B., Han X. Y., Sun, P., Yan, W. J. and Li, Y. C. (2011). Canonical correspondence analysis of algae community and its environmental factors in the Lake Chaohu, China. *Journal of Agro-Environment Science*, **30** (5), 952-958. (In Chinese)