Acute Effects of Cadmium and Copper on Cytochemical Responses in the Polychaete *Hediste diversicolor*

Bouraoui, Z. ^{1*}	, Ghedira, J	¹ , Banni	i, M. ^{1,2} and Boussetta,	H.1
----------------------------	--------------	----------------------	-------------------------------------	-----

¹Laboratory of Biochemistry and Environmental Toxicology, I.S.A. Chott-Mariem, 4042, University of Sousse, Tunisia

² Department of Environmental and Life Sciences, University of Piemonte Orientale Amedeo

Avogadro, 15100, Alessandria, Italy

Received 25 Jan. 2015;	Revised 30 June 2015;	Accepted 7 July 2015
------------------------	-----------------------	----------------------

ABSTRACT: This study investigated the cytochemical responses in *Hediste diversicolor* exposed to the environmental concentrations of cadmium (1 μ M Cd), copper (1 μ M Cu) and to their equimolar mixture for 48h. Lysosomal membrane stability (LMS), neutral lipids (NL) level, lipofuscin (LF) content and Ca²⁺-ATPase activity were determined as early subcellular biomarkers of toxicity. The biomarker battery pointed out a high cytoxicity on the polychaete treated with metals. Excessive induction of neutral lipids and lipofuscin accumulations were observed in polychaete treated with copper than cadmium. In addition, the highest level of neutral lipids and lipofuscin was noted in mixture treated group [Cu+Cd]. The increase of neutral lipids and lipofuscin accumulations were negatively correlated with the lysosomal membrane stability and Ca²⁺-ATPase activity alterations. Our study also demonstrated that the cytochemical biomarkers responses were very important to identify the cellular toxic effects of metals and promising biological tools in the assessment of biological effects of pollutants in natural marine environment.

Key word: Cadmium, Copper, Cytotoxicity, Cytochemical biomarkers, Hediste diversicolor

INTRODUCTION

Marine ecosystem is the final receptacle of various pollutants originating from human activities, urban, agricultural and industrial discharges. A considerable number of chemicals including metals have already been released into the environment and persist in sediment, water and biota. This situation endangers the health of organisms inhabiting this ecosystem (Levin et al., 2001; Wiegner et al., 2003). Metals have been of great concern in marine and coastal ecosystems, since they cause several biological alterations from molecular to tissue level depending on their concentrations and time exposure (Banni et al., 2007; Ben Khedher et al., 2014). Therefore, they may be also, accumulated in various tissues of living organisms and lead to several orders of magnitude higher than those of the surrounding water (Ghedira et al., 2011; Jebali et al., 2014).

Copper (Cu) is one of the most prevalent metals in coastal sediment (Salomons and Förstner, 1984) and at

high concentration, is highly toxic to marine life. Copper pollution in the aquatic environment results from natural and anthropogenic sources such as mine washing or agricultural leaching. Although copper is a trace element essential to life it is also one of the most toxic heavy metals (Tóth et al., 1996). Its action alters intracellular protein machinery either directly via denaturing enzymes or indirectly via generation of reactive oxygen species (Pourahmad and O'Brien, 2002; Droge, 2002) and has also recently been shown to cause DNA damage in sperm (Caldwell et al., 2011). Cadmium (Cd) is a heavy metal widely distributed in the environment as a result of natural and anthropogenic activities. Cadmium is commonly used in environmental studies because it is highly toxic (Lane and Morel, 2000) and can adversely affect the organisms at relatively low exposure concentrations (Waisberg et al., 2003). Cd is a toxic metal because of its interference with essential metals such as Zn and Fe, thereby interfering with many biochemical

^{*}Corresponding author E-mail: bouraoui_zied@yahoo.fr

processes that are regulated by metallic enzymes (Wright and Welbourn, 1994). In addition, Cd is also known to exert mutagenic (Mandel and Ryser, 1984) and carcinogenic (Degrave, 1981) effects, and induce lipid peroxidation and DNA breakage (Ochi and Ohsawa, 1983). Cd toxicity is also manifest as perturbations in metabolism such as enhanced lipolysis (Pierron *et al.*, 2007), and lowered glycogen content (Teh *et al.*, 1999; Bernsten and Lundebye, 2001). By contrast, very few data are available on other sublethal effects including lysosomal membrane stability, cytotoxicity, oxidative stress that could result from the biochemical activity of cadmium and copper in the polychaete *H. diversicolor*.

The polychaete Hediste diversicolor, is a marine annelid which lives in estuary sediments rich in microorganisms and toxic agents resulting from pollution. It has been the subject of numerous studies, focusing on different aspects of its biology and ecology, including a range of pollution related subjects. This polychaete is characterized by a high physiological tolerance to extreme variation of many environmental parameters such as temperature and salinity (Bartels-Hardege & Zeeck, 1990; Ait Alla et al., 2006). Also it's regarded as good bioindicator of metal and organic contamination in estuaries due to their highly tolerance of metal pollutants (Scaps, 2002; King et al., 2004; Bouraoui et al., 2010). A previous study found that the LC50 of Cd in H diversicolor is approximately 100 mg/L after 192 h of exposure (Bryan, 1976), which is much higher than that in most marine invertebrates (0.1-10 mg/L) (Cossa and Lassus, 1989). For copper, Yang et al. (2012) reported that the LC50 of Cu is approximately 475µg/L after 96 h of exposure. So it inferred that polychaetes have complex and effective detoxiûcation mechanisms to protect them from metal toxicity. Therefore, the use of polychaete worms as bioindicators for marine ecosystems has proved efficiency to be a useful tool in the assessment of environmental quality (Dean, 2008; Bouraoui et al., 2010).

Aquatic organisms are invariably exposed to pollutants mixtures. Studies examining bioaccumulation and subsequent toxicity of contaminants have often focused on single pollutants (Selck *et al.*, 1999; Pechenik *et al.*, 2001; Sandrini *et al.*, 2008). However, the interactions between these pollutants are poorly known and need to be studied. Considering the facts cited above, this work aimed to investigate the acute subcellular effects of copper and cadmium on the polychaeta *H.diversicolor*. A set of cellular stress markers was used including lysosomal membrane stability (LMS), lysosomal accumulation of lipofuscin (LF) and neutral lipids (NL); and plasma membrane Ca²⁺ ATPase activity.

MATERIALS & METHODS

Specimens of *H. diversicolor*, weighing between 0.4 and 0.6 g were collected in the intertidal zone at low tide from Teboulba (Tunisia) that was previously reported to be an unpolluted site (Banni *et al.*, 2005, 2007; Bouraoui *et al.*, 2009, 2014). The polychaetes and the surrounding sediments were placed in open polyethylene bottles and maintained at 14 °C to16 °C. Once in the laboratory the polychaetes were separated from sediment, cleaned from debris then placed in glass petri dishes at 14 °C with aerated clean seawater to ambient photoperiod regimes for 3 days. This acclimatation period was used for excluding specimens with exoskeleton or skin infections.

After this period, polychaetes were exposed for 48 h to the environmental concentration of 1 μ M of Cd (CdCl₂) or 1 μ M of Cu (CuSO₄) and to their equimolar mixture. A control group was run in parallel, employing only saline water (10‰) with the same characteristics cited above. After 48 h of exposure, worms washed brieûy in ice-cold buffer, fash-frozen in N-hexane, chilled in liquid nitrogen and then stored at -80 °C.

For Lysosomal membrane stability (LMS) assay, Pieces of excised worms were placed on two different aluminum cryostat chucks for a total of six individuals, and sections (10µM) were cut with a cryostat (Leica CM3050) and fash-dried by transferring them to room temperature. The cryostat sections were incubated at 37 °C in 50 mL of 0.1 M citrate buffer with 2.5% NaCl containing 20 mg naphthol AS-BI-N-acetyl- glucosaminide previously dissolved in 2.5 mL 2methoxyethanol and 3.5g of polypeptide, rinsed at 37 °C in 3% NaCl, treated at room temperature with 1 mg/ mL Fast Violet B in 0.1 M phosphate buffer and fixed for 15 min in Baker's fixative at 4 °C (Moore, 1976). Staining intensity of lysosomes was determined by studying the slide at 400×magnification with an inverted Axiovert microscope (Zeiss), connected to an Axiocam digital camera (Zeiss). Digital image analysis was carried out using the Scion Image software package (Scion Corp).

Neutral lipids content was made from sections prepared the same as for the lysosomal membrane stability assay by fixing the sections in calciumformaldehyde (2% Ca-acetate (w/v), 10% formaldehyde (v/v)) for 15 min at 4æ%C, followed by a rinsing step with de-ionised water, and incubation with 60% triethylphosphate (TEP) for 3 min. The sections were then stained with Oil Red-O (1% in 60% TEP) for 30 s, rinsed with de-ionised water, and mounted in 20% (v/ v) glycerol (Moore, 1988). Neutral lipid content was quantified by digital image analysis, as described for the LMS assay. For Lipofuscin content, Cryostat sections were fixed in calcium-formaldehyde and rinsed with deionised water, as described for the neutral lipid assay, followed by a 5 min incubation step with 1% Fe₂Cl₃, 1% potassium ferrocyanide in a 3:1 ratio (Moore, 1988). Te sections were rinsed with 1% acetic acid and mounted in 20% (v/v) glycerol. Lipofuscin content was quantifed by digital image analysis of stained sections, as described for the LMS assay.

The Ca²⁺-ATPase activity enzyme was quantified from cryostat sections obtained as for the LMS assay, using the histochemical method described by Pons et al., (2002). Cryostat sections were washed in 0.05M cacodylate buffer (pH 7.4), fixed in 1% paraformaldehyde (pFA) in 0.05 M cacodylate buffer (pH 7.4) for 30 min at 4 °C and washed again in 0.05M cacodylate buffer. Samples were then dehydrated in increasing acetone concentrations at 4°C and embedded in Technovit 7100 resin. Serial cross sections (2µm) were cut using a microtome, transferred onto glass slides and incubated for 6 h at room temperature in a medium containing 2.4 mM ATP, 18 mM CaCl., 8 mM levamisole, 0.2 mM ouabain, 1 mM Pb(NO₂)₂, and 20 mM sodium barbiturate. After incubation, the medium was removed and slides washed in water and rinsed in an ammonium sulfide-saturated water solution (3 min) to reveal the brown lead sulfide. $Pb_2(PO_4)_2$ precipitates stained with ammonium sulfide was quantified on sections by digital imaging as described above.

The results were expressed as means \pm SD. SPSS software (version 20.0) was used for statistical analysis. The data were first tested for normality and homogeneity of variance to meet statistical demands. Data from different groups were compared by a one-way analysis of variance (ANOVA). Gabriel's test was used to analyze raw biomarker data for comparison of responses between treatments. All differences were considered significant at p< 0.05. Different letters a, b, c and d indicated significant differences between groups.

RESULTS & DISCUSSION

No signiûcant worm mortality (<10%) was observed during the exposure period at different treatments with 1 μ M Cd, 1 μ M Cu or with their equimolar mixture of Cd and Cu. The effect of the tested single metal of Cu or Cd and their mixture on lysosomal membrane stability (LMS) is reported in Fig.1. A significant membrane destabilization with respect to untreated group was observed after exposure to each treatment with a percentage inhibition rising to 55% in copper group, 56% in cadmium group and 65% in mixture group. On the other hand, no significant differences (p<0.05) in terms of LMS was observed between treated groups. Representative images of intestinal cells staining of control and exposed animals are reported in Fig. 1.

The neutral lipid (NL) and lipofuscin (LF) contents in the treated *H. diversicolor* are reported in Fig 2 and 3. Lysosomal contents of NL and LF showed similar trend. In comparison to non exposed polychaetes, contents of NL signiûcantly increased in Cu-, Cd- and mixture- treated group.

The highest neutral lipid (NL) and lipofuscin (LF) accumulations levels were observed in (Cd+Cu) group (5.53-times higher than the control, P < 0.05). NL levels were 4.04-fold and 2.83-fold higher than the control group after single exposure respectively to Cu and Cd (Fig.2). When compared to control values, LF contents tended to increase in *H.diversicolor* exposed to contaminants (Fig. 3), but only the Cu and mixture exposure caused a significant increase in this parameter with more than respectively 435% and 1937%.

The Ca²⁺ ATPase activity, showed small changes between treatments tested with Cd and Cu (Fig. 4). However, the highest decrease in the activity was noted in worms exposed to the mixture with respect to the control reaching a percentage-value of 59%.

It has been accepted that aquatic ecosystems often contain mixtures of toxicants, including organic and inorganic compounds. It is important then to investigate the joint effects of more than one chemical pollutant on living organisms. In fact, individual components are likely to produce different responses within the same organisms and/or interact with other chemical pollutant producing additive, synergistic or antagonistic toxic effects (Barata et al., 2006). The influences of toxicant mixtures on natural communities are poorly understood and little is known about their interaction. However, the vast majority of available toxicity data, deals with the effects of single pure chemicals (Zhou, 1995). In the other hand, the joint effect of marine toxicants on the cytochemical markers in invertebrates was rarely investigated except for a few species (Koehler, 1992; Moore et al., 2004; Dondero et al. 2006 a,b; Gastaldi et al., 2007). In this work, we aimed to investigate the acute subcellular effects of copper and cadmium on the polychaeta H.diversicolor. A set of cellular stress markers were used including lysosomal membrane stability (LMS), lysosomal accumulation of lipofuscin (LF) and neutral lipids (NL); and plasma membrane Ca²⁺ ATPase activity.

The lysosomal system of many marine organisms including molluscs, annelids, crustaceans and ûsh, is known to be particularly sensitive to environmental perturbations and is generally thought to provide early



Fig 1. Acute effect of Cu, Cd and their mixture on LMS in *H. diversicolor* after 48 h exposure. Equal letters indicate absence of significant difference between means of different treatments. Representative images of cryostat tissue sections from control and exposed worms are reported. 1: epithelium, 2: coelome



Fig 2. Acute effect of Cu, Cd and their mixture on NL accumulation in *H. diversicolor* after 48 h exposure. Equal letters indicate absence of significant difference between means of different treatments. Representative images of cryostat tissue sections from control and exposed worms are reported. 1: coelome, 2: Neutral Lipid





Fig. 3. Acute effect of Cu, Cd and their mixture on lipofuscin (LF) content in *H. diversicolor* after 48 h exposure. Equal letters indicate absence of significant difference between means of different treatments. Representative images of cryostat tissue sections from control and exposed worms are reported. 1: epithelium, 2: lipofuscin



Fig. 4. Acute effect of Cu, Cd and their mixture on Ca²⁺ATPase activity in *H. diversicolor* after 48 h exposure. Equal letters indicate absence of significant difference between means of different treatments. Representative images of cryostat tissue sections from control and exposed worms are reported

response to pollutant exposure (Moore et al., 2006; Gastaldi et al., 2007, Bouraoui et al., 2014; Moschino et al., 2014). Lysosomal membrane stability (LMS) is a routinely used biomarker acting as an early indicator of the adverse effects of environmental contaminants both in invertebrates and vertebrates (Cajaraville et al., 2000; Galloway et al., 2004; Hankard et al., 2004; Gastaldi et al., 2006; Moore et al., 2006; Catalano et al., 2012). The effects on lysosomal membrane stability in H.diversicolor were evaluated in digestive cells following the reaction for N-acetyl- -hexosaminidase using histochemical procedures applied on frozen tissue sections. The obtained results demonstrated that both compounds (Cu 1 µM; Cd 1µM and [Cd+Cu] 1µM) induced a significant decrease in LMS after 48h of exposure underlining the high toxicity of these metals and confirming that lysosomes are a subcellular target for the action of both metal treatment. Similarly, previous studies in the polychaetes H. diversicolor including both laboratory exposure experiments and ûeld surveys reported a destabilization of lysosomal membrane under Cd, Cu and polycyclic aromatic hydrocarbons (PAHs) gradient (Catalano et al., 2012; Bouraoui et al., 2014). Interestingly, lysosomal membrane destabilisation was associated with a decrease of the Ca2+ ATPase activity in H. diversicolor exposed to Cd or/and Cu. Ca2+ ATPases are membrane SH-containing proteins and fundamental in the maintenance of cytosolic calcium homeostasis (Carafoli, 1994; Berridge et al., 1998). Marchi et al. (2004), have demonstrated that destabilisation of lysosomal membranes is linked to an increase in $[Ca^{2+}]$ and an activation of calcium-dependent phospholipase A2 (cPLA2). Also cPLA2 enzyme activity was the main factor for destabilisation of lysosomal membranes (Dondero et al., 2006). In other hand, several authors have demonstrated an inhibition of Ca2+ ATPase under effect of oxidative stress and heavy metals in both mammalian cells (Viarengo and Nicotera, 1991) and the cells of aquatic invertebrates such as mussels (Ermak and Davies, 2001; Burlando *et al.*, 2004), earthworms (Gastaldi *et al.*, 2007) and sea worms (Catalano *et al.*, 2012; Bouraoui *et al.*, 2014).

Toxic effects of metals are often depending on their capacity to increase the cellular levels of reactive oxygen species (ROS). When ROS levels production exceeds antioxidant defences, cells experience oxidative stress which causes, among others, membrane lipid peroxidation (Viarengo, 1989). Lipofuscins are end-products of lipid peroxidation which, due to their non-degradable nature, tend to accumulate in the lysosome vacuolar system. Our results showed that 1µM Cd, 1µM Cu and their equimolar mixture stimulate lipofuscinogenesis in H.diversicolor. It is well established that Cu is able to create an oxidative stress status by the Harber-Weiss and Fenton reaction of Cu cations (Suzuki et al., 1996), resulting in cellular damages due to hydroxyl radicals (HO•). For cadmium, many authors have already shown that Cd induce lipid peroxidation and DNA breakage (Ochi and Ohsawa, 1983; Bouraoui et al., 2009). Cd toxicity also manifest perturbations in metabolism such as enhanced lipolysis (Pierron et al., 2007) and decrease of adenylate energy charge (Den Besten et al., 1991). This holds even though it is well known that Cu is far more toxic than Cd in aquatic organisms such as ûsh, bivalves, and polychaete...(Cheng, 1990; Won et al., 2008).

Accumulation of neutral lipids (NL) in lysosomes is often associated with the effects of pollutants as a measure of unbalanced fatty acid metabolism (Catalano *et al.*, 2012). As reported by Moore (Moore, 1988), a build-up of these substances in digestive cells may be described as a form of lipidosis induced by toxic chemicals. In fact, our cytochemical data clearly indicate a stockpiling of neutral lipids in the cytoplasm of cells of organisms exposed to 1μ M Cd, 1μ M Cu and then peaked in those treated with equimolar mixture. Data reported in the present work show a strong and constant response of the cytochemical parameters in worms exposed either to only Cu/Cd or to the mixture (Cu+ Cd). Our results have highlighted an additive effect of the binary mixtures (Cu+Cd) as already observed in previous studies at similar concentrations and time on enzymatic biomarkers and lipoperoxidation level in *H. diversicolor*. (Banni *et al.*, 2009; Bouraoui *et al.*, 2009).

CONCLUSIONS

This work demonstrates clearly, the subcellular negative effects of the environmental concentration of copper and cadmium in the polychaete in *H. diversicolor*. Remarkable alterations of the lysosomal membrane stability, lipid peroxidation and Ca²⁺ATPase activity were observed. Excessive induction of neutral lipids and lipofuscin accumulations were observed in polychaete treated with copper than cadmium. The increase of neutral lipids and lipofuscin accumulations were negatively correlated with the lysosomal membrane stability and Ca²⁺-ATPase activity alterations. In addition, our results indicate that the polychaete *H.diversicolor* should be a valuable benthic organism for studying risks of marine pollutants using cytochemical tests.

ACKNOWLEDGEMENTS

This work was conducted within the framework of MOBIDOC-Post doc as specified in PASRI program, funded by the EU and managed by the ANPR. It was partially supported by the Research Unit of Biochemistry and Environmental Toxicology, UR 04AGR05) (IRESA, Tunisia).

REFERENCES

Ait Alla, A., Mouneyrac, C., Durou, C., Moukrim, A. and Pellerin, J. (2006). Tolerance and biomarkers as useful tools for assessing environmental quality in the Oued Souss estuary (Bay of Agadir, Morocco). Comparative Biochemistry and Physiology - Part C, **143**(1), 23-29.

Banni, M., Jebali, J., Daubeze, M., Clerendeau, C., Guerbej, H., Narbonne, J. F. and Boussett, H. (2005). Monitoring Pollution in Tunisian coasts: Application of a classification scale based on biochemical markers. Biomarkers, 10, 105-116.

Banni, M., Dondero, F., Jebali, J., Guerbej, H., Boussetta, H. and Viarengo, A. (2007). Assessment of heavy metal contamination using real time PCR analysis of mussel metallothionein mt10 and mt20 expression: a validation along the Tunisian coasts. Biomarkers, **12(4)**, 369-383.

Banni, M., Bouraoui, Z., Clerandeau, C, Narbonne, J.F. and Boussetta, H. (2009). Mixture toxicity assessment of cadmium and benzo[a]pyrene in the sea worm *Hediste diversicolor*. Chemosphere, **77**(**7**), 902-906.

Barata, C., Baird, D.J., Nogueira, A.J.A., Soares, A.M. and Riva, M.C. (2006). Toxicity of binary mixtures of metals and pyrethroid insecticides to Daphnia magna Straus. Implications for multi-substance risks assessment. Aquatic Toxicology, **78(1)**, 1-14.

Bartels-Hardege, H.D. and Zeeck, E. (1990). Reproductive behaviour of *Nereis diversicolor* (Annelida: Polychaeta). Marine Biology, **106**, 409-412.

Ben-Khedher, S., Jebali, J., Houas, Z., Nawéli, H., Jrad, A., Banni, M. and Boussetta, H. (2014). Metals bioaccumulation and histopathological biomarkers in *Carcinus maenas* crab from Bizerta lagoon, Tunisia. Environmental science and pollution research international, **21(6)**, 4343-4357.

Bernsten, M. and Lundebye, A.K.H. (2001). Energetics in Atlantic salmon (Salmo salar L.) parr fed elevated dietary cadmium. Comparative Biochemistry and Physiology - Part C, **128(3)**, 311-323.

Berridge, M.J. and Bootman, M.D. and Lipp, P. (1998). Calcium a life and death signal. Nature **395**, 645–648.

Bouraoui, Z., Banni, M., Ghedira, J., Clerandeau, C., Narbonne, J.F. and Boussetta, H. (2009). Evaluation of enzymatic biomarkers and lipoperoxidation level in *Hediste diversicolor* exposed to copper and benzo[a]pyrene. Ecotoxicology and Environmental Safety, **72**(7), 1893-1898.

Bouraoui, Z., Banni, M., Chouba, L., Ghedira, J., Clerandeau, C., Jebali, J., Narbonne, J.F., Boussetta, H. (2010). Monitoring pollution in Tunisian coasts using a scale of classification based on biochemical markers in worm *Nereis (Hediste) diversicolor*. Environmental Monitoring and Assessment, **164(1-4)**, 691-700.

Bouraoui, Z., Ghedira, Capri, F., Chouba, L. and Boussetta, H. (2014). Cytochemical responses of *Hediste diversicolor* (Nereidae, Polychaete) sampled from polluted sites along the Tunisian coast. Journal of Integrated Coastal Zone Management, **14(1)**, 119-127.

Bryan, G.W. (1976). Heavy metal contamination in the sea. In: Johnston, R. (Ed.), Marine Pollution. London Academic Press, London, pp. 185-302.

Burlando, B., Bonomo, M., Capri, F., Mancinelli, G., Pons, G., Viarengo, A. (2004). Different effects of Hg^{2+} and Cu^{2+} on mussel (Mytilus galloprovincialis) plasma membrane Ca^{2+} -ATPase: Hg^{2+} induction of protein expression. Comparative Biochemistry and Physiology - Part C, **139(4)**, 201-207.

Cajaraville, M.P., Bebianno, M.J., Blasco, J., Porte, C., Sarasquete, C. and Viarengo, A. (2000). The use of biomarkers to assess the impact of the Iberian Peninsula: a pratical approach. Science of the Total Environment, **247(2-3)**, 295-311.

Caldwell, G.S., Lewis, C., Pickavance, G., Taylor, R.L. and Bentley, M.G. (2011). Exposure to copper and cytotoxic polyunsaturated aldehyde induces reproductive failure in the marine polychaete Nereis virens (Sars). Aquatic Toxicology, **104(1-2)**, 126-134.

Carafoli, E. (1994). The signaling function of calcium and its regulation. Official Journal of the International Society of Hypertension, **12(10)**, 547-556.

Catalano, B., Moltedo, G., Martuccio, G., Gastaldi, L., Virno-Lamberti, C., Lauria, L. and Ausili, A. (2012). Can *Hediste diversicolor* (Nereidae, Polychaete) be considered a good candidate in evaluating PAH contamination? A multimarker approach. Chemosphere, **86(9)**, 875-882.

Cheng, T.C. (1990). Effects of *in vivo* exposure of *Crassostrea virginica* to heavy metals on hemocyte viability and activity levels of lysosomal enzymes. In: Perkins, O.F., Cheng, T.C. (Eds.), Pathology in Marine Science. Academic Press, London, pp. 513–524.

Dean, H.K. (2008). The use of polychaetes (Annelida) as indicator species of marine pollution: a review. International Journal of Tropical Biology and Conservation, **56(4)**, 11-38.

Degrave, N. (1981). Carcinogenic, teratogenic and mutagenic effects of cadmium. Mutation Research, **86(1)**, 115-135.

Den Besten, P.J., Bosma, P.T., Herwig, H.J., Zandee, D.I. and Voogt, P.A. (1991). Effects of cadmium on metal composition and adenylate energy charge in the sea star *Asterias rubens L*. Archives of Environmental Contamination and Toxicology, **21**(1), 112-117.

Dondero, F., Dagnino, A., Jonsson, H., Capri, F., Gastaldi, L. and Viarengo, A. (2006 a). Assessing the occurrence of a stress syndrome in mussels (Mytilus edulis) using a combined biomarker/gene expression approach. Aquatic Toxicology, **78**(1), 13-24.

Dondero, F., Jonsson, H., Rebelo, M., Pesce, G., Berti, E., Pons, G. and Viarengo, A. (2006 b). Cellular responses to environmental contaminants in amoebic cells of the slime mould *Dictyostelium discoideum*. Comparative Biochemistry and Physiology - Part C, **143(2)**, 150-157.

Droge, W. (2002). Free radicals in the physiological control of cell function. Physiological Reviews, **82(1)**, 47-95.

Ermak, G. and Davies, K.J. (2001). Calcium and oxidative stress: from cell signaling to cell death. Molecular Immunology, **38(10)**, 713-721.

Galloway, T.S., Brown, R.J., Browne, M.A., Dissanayake, A., Lowe, D. and Jones, M.B. (2004). A multibiomarker approach to environmental assessment. Environmental Science & Technology, **38(6)**, 1723-1731.

Gastaldi, L., Ranzato, E., Capri, F., Hankard, P., Peres, G., Canesi, L., Viarengo, A. and Pons, G., (2007). Application of a biomarker battery for the evaluation of the sublethal effects of pollutants in the earthworm *Eisenia andrei*. Comparative Biochemistry and Physiology - Part C, **146(3)**, 398-405.

Ghedira, J., Jebali, J., Banni, M., Chouba, L., Boussetta, H., López-Barea, J. and Alhama, J. (2011). Use of oxidative stress biomarker sin *Carcinus maenas* to assess littoral zone contamination in Tunisia. Aquatic Biology, **14**, 87-98. Hankard, P.K., Svendsen, C., Wright, J., Wienberg, C., Fishwick, S.K. and Spurgeon, D.J. (2004). Biological assessment of contaminated land using earthworm biomarkers in support of chemical analysis. Science of the Total Environment, **330(1-3)**, 9-20.

Jebali, J., Chicano-Gálvez, E., Fernández-Cisnal, R., Banni, M., Chouba, L., Boussetta, H., Barea, H.L. and Alhama, J. (2014). Proteomic analysis in caged Mediterranean crab (Carcinus maenas) and chemical contaminant exposure in Téboulba Harbour, Tunisia. Ecotoxicology and Environmental Safety, **100**, 15-26.

King, C.K., Dowse, M.C., Simpson, S.L. and Jolley, D.F. (2004). An assessment of five Australian polychaetes and bivalves for use in whole-sediment toxicity tests: toxicity and accumulation of copper and zinc from water and sediment. Archives of Environmental Contamination and Toxicology, **47**(3), 314-323.

Koehler, A., Deisemann, H. and Lauritzen, B. (1992). Histological and cytochemical indices of toxic injury in the liver of the dab Limanda limanda. Marine Ecology Progress Series, **91**, 141-153.

Lane, T.W. and Morel, F.M.M. (2000). A biological function for cadmium in marine diatoms. Proceedings of the National Academy of Sciences of the United States of America, **97(9)**, 4627-4631.

Levin, L.A., Boesch, D.F., Covich, A., Dahm, C., Erséus, C., Ewel, K.C., Kneib, R.T., Moldenke, A., Palmer, M.A., Snelgrove, P., Strayer, D. and Weslawski, J.M. (2001). The function of marine critical transition zones and the importance of sediment biodiversity. Ecosystems, **4**, 430-451.

Mandel, R. and Ryser, H.J.P. (1984). Mutagenicity of cadmium in Salmonella typhimurium and its synergism with two nitrosamines. Mutation Research, **138**, 9-16.

Marchi, B., Burlando, B., Moore, M.N. and Viarengo, A. (2004). Mercury and copper induced lysosomal membrane destabilisation depends on $[Ca^{2+}]_i$ dependent phospholipase A2 activation. Aquatic Toxicology, **66(2)**, 197-204.

Moore, M.N. (1976). Cytochemical demonstration of latency of lysosomal hydrolases in digestive gland cells of the common mussel Mytilus edulis, and changes induced by thermal stress. Cell and Tissue Research, **175(3)**, 279-287.

Moore, M.N. (1988). Cytochemical responses of the lysosomal system and NADPH-ferrihemoprotein reductase in molluscan digestive cells to environmental and experimental exposure xenobiotics. Marine Ecology Progress Series, **46(1-3)**, 81-89.

Moore, M.N., Depledge, M.H., Readman, J.W. and Leonard, D.R.P. (2004). An integrated biomarker-based strategy for ecotoxicological evaluation of risk in environmental management. Mutation Research, **552**(1-2), 247-268.

Moore, M.N., Icarus, A.J., McVeigh, A. (2006). Environmental prognostics: an integrated model supporting lysosomal stress responses as predictive biomarkers of animal health status. Marine Environmental Research, **61(3)**, 278-304.

Moschino, V., Nestoa, N., Barisonb, S., Agresti, F., Colla, L., Fedele, L. and Da Ros, L. (2014). A preliminary investigation on nanohorn toxicity in marine mussels and polychaetes. Science of The Total Environment, 468-469, 111-119.

Ochi, T. and Ohsawa, M. (1983). Induction of 6-thioguanineresistant mutants and single strand scission of DNA by cadmium chloride in cultured Chinese hamster cells. Mutation Research, **111(1)**, 69-78.

Pechenik, J.A., Berard, R., Daniels, D., Gleason, T.R. and Champlin, D. (2001). Influence of lowered salinity and elevated cadmium on the survival and metamorphosis of trochophores in *Capitella sp.* I. Invertebrate Biology, **120(2)**, 142-148.

Pierron, F., Baudrimont, M., Bossy, A., Bourdineaud, J.-P., Brèthes, D., Elie, P. and Massabuau, J.C. (2007). Impairment of lipid storage by cadmium in the European eel (Anguilla anguilla). Aquatic Toxicology, **81**(**3**), 304-311.

Pons, G., Evangelisti, V., Capri, F., Mozzone, S. and Viarengo, A. (2002). Cytochemical localization and quantification of plasma membrane Ca^{2+} - ATPase activity in mollusk digestive gland cells. European Journal of Histochemistry, **46(1)**, 31-40.

Pourahmad, J. and O'Brien, P.J. (2002). A comparison of hepatocyte cytotoxic mechanisms for Cu^{2+} and Cd^{2+} . Toxicology, **143(3)**, 263–273.

Sandrini, J.Z., Ventura Lima, J., Regoli, F., Fattorini, D., Notti, A., Marins, L.F. and Monserrat, J.M., (2008). Antioxidant responses in the nereidid *Laeonereis acuta* (Annelida, Polychaeta) after cadmium exposure. Ecotoxicology and Environmental Safety, **70(1)**, 115-120.

Scaps, P. (2002). A review of the biology, ecology and potential use of the common ragworm *Hediste diversicolor* (O.F. Muller) (Annelida: Polychaeta). Hydrobiologia, **470**(1-**3**), 203-218.

Salomons, W. and Förstner, U. (1984). Metals in the Hydrocycle. Springer-Verlag, Berlin, pp. 349.

Selck, H., Decho, A.W. and Forbes, V.E. (1999). Effects of chronic metal exposure and sediment organic matter on digestive absorption efficiency of cadmium by the deposit-feeding polychaete *Capitella* species I. Environmental Toxicology and Chemistry. **18(6)**, 1289-1297.

Teh, S.J., Clark, S.L., Brown, C.L., Luoma, S.N. and Hinton, D.E. (1999). Enzymatic and histopathologic biomarkers as indicators of contaminant exposure and effect in Asian clam (Potamocorbula amurensis). Biomarkers, **4(6)**, 497-509.

Toth, L., Juhàsz, M., Varga, T., Csikkel-Szolnoki, A. and Nemcsok, J. (1996). Some effect of $CuSO_4$ on carp. Journal of Environmental Science and Health, Part B, **31(3)**, 627-635.

Viarengo, A. (1989). Heavy metals in marine invertebrates: mechanisms of regulation and toxicity at the cellular level. Aquatic Sciences, **1**, 295-317. Viarengo, A. and Nicotera, P. (1991). Possible role of Ca^{2+} in heavy metal cytotoxicity. Comparative Biochemistry and Physiology Part C, **100(1-2)**, 81-84.

Waisberg, M., Joseph, P., Hale, B. and Beyersmann, D. (2003). Molecular and cellular mechanisms of cadmium carcinogenesis. Toxicology, **192(2-3)**, 95-117.

Wiegner, T.N., Seitzinger, S.P., Breitburg, D.L. and Sanders, J.G. (2003). The effects of multiple stressors on the balance between autotrophic and heterotrophic processes in an estuarine system. Estuaries, **26(2)**, 352-364.

Wright, D.A. and Welbourn, P.M. (1994). Cadmium in the aquatic environment: a review of ecological, physiological, and toxicological effects on biota. Environmental Reviews. **2(2)**, 187-214.

Won, E.J., Raisuddin, S. and Shin, K.H. (2008). Evaluation of induction of metallothionein like proteins (MTLPs) in the polychaetes for biomonitoring of heavy metal pollution in marine sediment. Marine Pollution Bulletin, **57(6-12)**, 544-551.

Yang, D., Zhoua, Y., Zhao, H., Zhoua, X., Sun N., Wang, B. and Yuan, X. (2012). Molecular cloning, sequencing, and expression analysis of cDNA encoding metalloprotein II (MP II) induced by single and combined metals (Cu(II), Cd(II)) in polychaeta *Perinereis aibuhitensis*. Environmental Toxicology and Pharmacology, **34**(**3**), 841-848.

Zhou, J.L., Rowland, S.J., Mantoura, R.F.C. and Harland, B.J. (1995). Influence of the nature of particulate organic matter on the sorption of cypermethrin: implications on K_{oc} correlations. Environment International, **21**(2), 187-195.